Total Maximum Daily Load of Sediment in the Non-Tidal Upper Chester River Watershed, Kent and Queen Anne's Counties, Maryland

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Upper Chester River
Sediment TMDL
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List of Abbreviations

AFO	Animal Feeding Operations
ANC	Acid Neutralizing Capacity
BIBI	Benthic Index of Biotic Integrity
BIP	Buffer Incentive Program
BMP	Best Management Practices
BSID	Biological Stressor Identification
CAFOs	Concentrated Animal Feeding Operations
CBLCD	Chesapeake Bay Land-Cover Dataset
CBP	Chesapeake Bay Program
CBP P4.3	Chesapeake Bay Program Model Phase 4.3
CBP P5.3.2	Chesapeake Bay Program Model Phase 5.3.2
CBT	Chesapeake Bay Trust
CCAP	Coastal Change Analysis Program
CFR	Code of Federal Regulations
cfs	Cubic Feet per Second
CHSTF	Upper Chester River Tidal Fresh
COMAR	Code of Maryland Regulations
CV	Coefficient of Variation
CWA	Clean Water Act
DI	Diversity Index
EOF	Edge-of-Field
EOS	Edge-of-Stream
EPT	*Ephemeroptera*, *Plecoptera*, and *Trichoptera*
ESD	Environmental Site Design
FIBI	Fish Index of Biologic Integrity
GIS	Geographic Information System
HBI	Hilsenhoff Biotic Index
HSPF	Hydrological Simulation Program Fortran
IBI	Index of Biotic Integrity
LA	Load Allocation
m	Meter
m^3/yr	Meters cubed per year
MACS	Maryland Agricultural Water Quality Cost-Share Program
MAL	Minimum Allowable IBI Limit
MBSS	Maryland Biological Stream Survey
MDA	Maryland Department of Agriculture
MDDNR	Maryland Department of Natural Resources
MDE	Maryland Department of the Environment
MDL	Maximum Daily Load
MDP	Maryland Department of Planning
MGD	Millions of Gallons per Day
mg/l	Milligrams per liter

MGS	Maryland Geological Survey
MOS	Margin of Safety
MRLC	Multi-Resolution Land Characteristics
MS4	Municipal Separate Storm Sewer System
NLCD	National Land-Cover Dataset
NOAA	National Oceanic and Atmospheric Administration
NPDES	National Pollutant Discharge Elimination System
NRCS	Natural Resource Conservation Service
NRI	Natural Resources Inventory
PSU	Primary Sampling Unit
SCS	Soil Conservation Service
SDF	Sediment Delivery Factor
SHA	State Highway Administration
TMDL	Total Maximum Daily Load
ton/acre/yr	Tons per acre per year
ton/day	Tons per day
ton/yr	Tons per year
TSD	Technical Support Document
TSS	Total Suspended Solids
USDA	United States Department of Agriculture
USEPA	U.S. Environmental Protection Agency
USGS	United States Geological Survey
WIP	Watershed Implementation Plan
WLA	Waste Load Allocation
WQA	Water Quality Analysis
WQLS	Water Quality Limited Segment

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EXECUTIVE SUMMARY

Section 303(d) of the federal Clean Water Act (CWA) and the U.S. Environmental Protection Agency's (USEPA) implementing regulations direct each state to identify and list waters, known as water quality limited segments (WQLSs), in which current required controls of a specified substance are inadequate to achieve water quality standards. For each WQLS, the State is required to either establish a TMDL of the specified substance that the waterbody can receive without violating water quality standards, or demonstrate that water quality standards are being met (CFR 2012b). This document, upon approval by USEPA, establishes a Total Maximum Daily Load (TMDL) for sediment/total suspended solids (TSS) in the non-tidal Maryland 8-Digit Upper Chester River watershed (2016 *Integrated Report of Surface Water Quality in Maryland* Assessment Unit ID: MD-02130510). In this TMDL report, the terms total suspended solids (TSS) and sediment may be used interchangeably.

The Upper Chester River watershed is associated with two assessment units in Maryland's Integrated Report: a non-tidal 8-digit watershed (02130510) and an estuary portion [Upper Chester River Tidal Fresh (CHSTF)]. This TMDL only addresses sediment loads in the non-tidal portion of the watershed. Upstream loads from the non-tidal Delaware Chester River discharge into the non-tidal 8-digit watershed and are included in this TMDL. A sediment TMDL for the CHSTF was established as part of the Chesapeake Bay TMDLs in 2010. Background information on the tidal portion of the watershed is presented for informational purposes only. Only a baseline load and load allocation have been calculated for the Delaware Chester River.

The Maryland Department of the Environment (MDE) identified the waters of the Upper Chester River watershed and associated assessment units on the State's 2016 Integrated Report as impaired by multiple pollutants (MDE 2016). Table ES-1 identifies Integrated Report listings associated with this watershed. The upstream Delaware Chester River is not listed by Delaware as impaired by sediment, but is prescribed sediment reductions in this TMDL at the State boundary in order to meet the Maryland TMDL. A data solicitation for sediment was conducted by MDE in March 2018, and all readily available data has been considered.

The Maryland Surface Water Use Designation in the Code of Maryland Regulations (COMAR) for the Upper Chester River watershed's non-tidal tributaries are designated as Use Class I – *water contact recreation and protection of aquatic life.* Tidal tributaries and the Upper Chester River mainstem are designated Use Class II – *support of estuarine and marine aquatic life and shellfish harvesting* (COMAR 2016a, b, c).

The non-tidal Upper Chester River watershed was originally listed for biological impairment on the 2006 Integrated Report. The listing was based on the biological assessment methodology, which uses aquatic health scores, consisting of the Benthic Index of Biotic Integrity (BIBI) and Fish Index of Biotic Integrity (FIBI). These indices indicated that the biological metrics for the watershed exhibit a significant negative deviation from reference conditions (MDE 2006).

Table ES-1: Upper Chester River 2016 Integrated Report Listings

Watershed	Basin Code	Tidal/Non-tidal	Designated Use Class	Year Listed	Identified Pollutant	Listing Category
Upper Chester River	02130510	Non-tidal	I - Aquatic Life and Wildlife	2012	TSS	5
					Channelization	4c
Millington Wildlife Ponds		Impoundment	Fishing	2004	Mercury in Fish Tissue	4a
Upper Chester River Tidal Fresh	CHSTF	Tidal	II – Seasonal migratory fish spawning and nursery	1996	TN	4a
				1996	TP	4a
			II – Aquatic Life and Wildlife	1996	Impacts to Estuarine Biological Communities	3
			II – Open Water Fish and Shellfish	1996	TP	4a
					TN	4a
			II – Seasonal Shallow-Water Submerged Aquatic Vegetation Subcategory	-	TSS	2
Duck Neck Beach			II – Water Contact Sports	1996	Enterococci	4a

Note:
- Category 2 indicates the waterbody is meeting water quality standards for the identified substance
- Category 3 indicates insufficient data to make a listing category determination
- Category 4a indicates a TMDL has been completed and approved by EPA
- Category 4c indicates the cause of the impairment is pollution and not a pollutant
- Category 5 indicates that the waterbody is impaired and a TMDL or water quality analysis (WQA) is needed.

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In order to determine what stressor or stressors are impacting aquatic life, MDE's *Biological Stressor Identification* (BSID) methodology was applied. The BSID analysis for the Upper Chester River watershed identified sediment and instream habitat parameters. The sediment parameter group shows a significant association with poor epifaunal substrate. The instream habitat parameter group shows a significant association with channelization present, marginal to poor instream habitat structure, poor riffle/run quality, poor velocity/depth diversity, and concrete/gabion present. Further details of this analysis are presented in the 2012 document entitled, *Watershed Report for Biological Impairment of the Non-Tidal Upper Chester River Watershed in Kent and Queen Anne's Counties Counties, Maryland Biological Stressor Identification Analysis Results and Interpretation* (MDE 2012).

As a result of the BSID analysis, the non-tidal MD 8-digit Upper Chester River watershed was listed on the 2012 Integrated Report as impaired by TSS thus requiring a TMDL. The TMDL will apply only to the non-tidal portion of the watershed. For simplicity, further reference in this document to Upper Chester River Watershed will refer only to the non-tidal MD 8-digit watershed.

The objective of this TMDL is to ensure that watershed sediment loads are at a level that supports the Use Class I designations for the non-tidal Upper Chester River watershed. The TMDL will address impacts to aquatic life in the non-tidal Upper Chester River watershed caused by high sediment and TSS concentrations. Separate sediment TMDLs were developed for the Use Class II impairments as part of the Chesapeake Bay TMDLs in 2010.

The CWA requires TMDLs to be protective of all the designated uses applicable to a particular waterbody. The primary focus of this TMDL is the designated use of protection of aquatic life because the Integrated Report listing was based on a biological assessment of the watershed. The biological assessment revealed the current levels of TSS and other pollutants prevent the watershed from achieving its designated use of supporting aquatic life. The required reductions within the TMDL are expected to protect all designated uses of the watershed from sediment impacts, including water contact recreation. Aquatic life is more sensitive to sediment impacts than recreation because of continuous exposure that can affect respiration and propagation. Recreation, on the other hand, is sporadic and sediment is unlikely to pose a human health risk due to dermal contact or minimal ingestion that would occur during recreation.

Currently in Maryland, there are no specific numeric criteria that quantify the impact of sediment on the aquatic life of non-tidal stream systems. In order to quantify this impact, a reference watershed TMDL approach was used, which resulted in the establishment of a *sediment loading threshold* (MDE 2006). This threshold is based on a detailed analysis of sediment loads from watersheds that are identified as supporting aquatic life (i.e., reference watersheds) based on Maryland's biological assessment methodology (Roth et al. 1998, 2000; Stribling et al. 1998; MDE 2014b). This threshold is then used to determine watershed specific sediment TMDL endpoint. The resulting loads are

considered the maximum allowable loads the waterbody can receive without causing any sediment related impacts to aquatic health.

In order to use a reference watershed approach, sediment loads are estimated using a watershed model. For this analysis, the Chesapeake Bay Program Phase 5.3.2 (CBP P5.3.2) watershed model was chosen and specifically, the *edge-of-stream* (EOS) land-use sediment loads were used. The CBP P5.3.2 model was appropriate for this TMDL because the spatial domain of the model segmentation aggregates to the MD 8-digit watershed scale, which is consistent with the impairment listing.

USEPA's regulations require TMDLs to take into account seasonality and critical conditions for stream flow, loading, and water quality parameters (CFR 2012b). The intent of this requirement is to ensure that the water quality of the waterbody is protected during times when it is most vulnerable. The biological monitoring data used to determine the reference watersheds reflect the impacts of stressors (i.e., sediment impacts to stream biota) over the course of time (i.e., captures the impacts of both high and low flow events). Thus, critical conditions are inherently addressed. Seasonality is captured in several components. First, it is implicitly included in biological sampling as biological communities reflect the impacts of stressors over time, as described above. Second, the Maryland Biological Stream Survey (MBSS) dataset, which serves as the primary dataset for calculating the biological metrics of the watershed (i.e., BIBI and FIBI scores), included benthic sampling in the spring and fish sampling in the summer. Moreover, the sediment loading rates used in the TMDL were determined using the CBP P5.3.2 model, which is a continuous simulation model with a simulation period 1991-2000, based on Hydrological Simulation Program Fortran (HSPF) model, thereby addressing annual changes in hydrology and capturing wet, average, and dry years.

All TMDLs need to be presented as a sum of waste load allocations (WLAs) for point sources and load allocations (LAs) for nonpoint sources generated within the assessment unit, accounting for natural background, tributary and adjacent segment loads. Furthermore, all TMDLs must include a margin of safety (MOS) to account for any lack of knowledge and uncertainty concerning the relationship between loads and water quality (CFR 2012a,b). It is proposed that the estimated variability around the reference watershed group used in this analysis already accounts for such uncertainty, and therefore the MOS is implicitly included. Because the sediment loading threshold was conservatively based on the median (50th percentile) sediment loading rates from reference watersheds, Maryland has adopted an implicit MOS for sediment TMDLs.

The Upper Chester River watershed total baseline sediment load is 7,150 tons per year (ton/yr). This baseline load consists of upstream loads generated outside the assessment unit (i.e., MD 8-digit watershed): a Delaware Chester River Baseline Load (BL_{DE}) of 1,170 ton/yr, and loads generated within the assessment unit: an Upper Chester River Watershed Baseline Load Contribution of 5,980 ton/yr. The Upper Chester River watershed baseline load contribution is further subdivided into a nonpoint source baseline load (Nonpoint Source BL_{UCR}) and two types of point source baseline loads: National

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Pollutant Discharge Elimination System (NPDES) regulated stormwater (NPDES Stormwater BL$_{UCR}$) and NPDES regulated wastewater (Wastewater BL$_{UCR}$) (see Table ES-2).

Table ES-2: Upper Chester River Watershed Baseline Sediment Loads (ton/yr)

		Upstream Baseline Load[1]		Upper Chester River Watershed Baseline Load Contribution				
Total Baseline Load	=	BL$_{DE}$	+	Nonpoint Source BL$_{UCR}$	+	NPDES Stormwater BL$_{UCR}$	+	Wastewater BL$_{UCR}$
7,150	=	1,170	+	5,955	+	15	+	10

The Upper Chester River Watershed average annual TMDL of TSS is 5,673 ton/yr (a 21% reduction from the baseline load). The TMDL consists of allocations attributed to loads generated outside the assessment unit referred to as Upstream Load Allocations: a Delaware Chester River Load Allocation of 916 ton/yr (LA$_{DE}$); and loads generated within the assessment unit: an Upper Chester River Watershed TMDL Contribution of 4,757 ton/yr. The Upper Chester River TMDL contribution is further subdivided into point and nonpoint source allocations and is comprised of a load allocation (LA$_{UCR}$) of 4,732 ton/yr, an NPDES Stormwater Waste Load Allocation (NPDES Stormwater WLA$_{UCR}$) of 15 ton/yr, and a Wastewater Load Allocation (Wastewater WLA$_{UCR}$) of 10 ton/yr (see Table ES-3).

Table ES-3: Upper Chester River Watershed Average Annual TMDL of Sediment (ton/yr)

TMDL	=	LA			+	LA$_{UCR}$	+	WLA			+	MOS
		LA$_{DE}$	+					NPDES Stormwater WLA$_{UCR}$	+	Wastewater WLA$_{UCR}$	+	
5,673	=	916	+			4,732	+	15	+	10	+	Implicit

Upstream Load Allocations[,1]   Upper Chester River Watershed TMDL Contribution

Notes: [1] A delivery factor of 1 was used for all of the Upstream Load Allocations.

Table ES-4: Upper Chester River Watershed Baseline Load, TMDL, and Total Reduction Percentage

Baseline Load (ton/yr)	TMDL (ton/yr)	Total Reduction (%)
7,150	5,673	21

Note: The load summary includes the Delaware Chester River Upstream Baseline Load and TMDL Load Allocation.

In addition to the TMDL value, a Maximum Daily Load (MDL) is also presented in this document. The calculation of the MDL, which is derived from the TMDL average annual loads, is explained in Appendix B and presented in Table B-1.

This TMDL will ensure that watershed sediment loads are at a level to support the Use Class I designation for the non-tidal Upper Chester River watershed, and more specifically, at a level to support aquatic life. The TMDL will not completely resolve the impairment to biological communities within the watershed since the BSID watershed analysis identifies other possible stressors impacting the biological conditions (e.g. channelization).

Section 303(d) of the CWA and current USEPA regulations require reasonable assurance that the TMDL can and will be implemented. Once USEPA has approved this TMDL and it is known what measures must be taken to reduce pollution levels, implementation of best management practices (BMPs) is expected to take place. MDE intends for the required TMDL reductions to be implemented in an iterative process that first addresses those sources with the largest impact to water quality, with consideration given to cost of implementation.

Implementation of the Upper Chester River Watershed Sediment TMDL is expected to occur in conjunction with implementation efforts to meet sediment target loads consistent with the Chesapeake Bay TMDLs. The Chesapeake Bay TMDLs were established by USEPA and are scheduled for full implementation by 2025 (USEPA 2010a). These TMDLs require reductions of nitrogen, phosphorus, and sediment loads throughout the Bay watershed to meet water quality standards that protect the designated uses in the Bay and its tidal tributaries.

While this TMDL establishes a sediment loading target for the watershed, watershed managers and other stakeholders should always remain cognizant that the endpoint of this TMDL, and hence the definition of its successful implementation, is based on in-stream biological health. Load reductions are critical to tracking this effort, since the TMDL target is defined as the point where sediment loads match those seen in reference watersheds, but the watershed cannot be delisted or classified as meeting water quality standards until it is demonstrated that the biological health of the stream system is no longer impaired by sediment. In planning any implementation efforts related to this TMDL, careful consideration should be given both to the sediment load reductions, and to their direct potential impacts on biological communities.

Many practices in the implementation plans that reduce sediment concurrently address other stressors identified in the BSID report. Since biological improvements will likely only be seen when multiple structural and pollutant stressors are addressed, watershed managers developing plans to address sediment should consider the effect of restoration projects on other stressors. Where possible, preference should be given to designs that address multiple stressors.

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1.0 INTRODUCTION

Section 303(d) of the federal Clean Water Act (CWA) and the U.S. Environmental
Protection Agency's (USEPA) implementing regulations direct each state to identify and
list waters, known as water quality limited segments (WQLSs), in which current required
controls of a specified substance are inadequate to achieve water quality standards. For
each WQLS, the State is required to either establish a TMDL of the specified substance
that the waterbody can receive without violating water quality standards, or demonstrate
that water quality standards are being met (CFR 2012b). This document, upon approval
by the USEPA, establishes a Total Maximum Daily Load (TMDL) for sediment in the
non-tidal Maryland 8-Digit Upper Chester River watershed (2016 *Integrated Report of
Surface Water Quality in Maryland* Assessment Unit ID: MD-02130510). In this TMDL
report, the terms total suspended solids (TSS) and sediment may be used interchangeably.

TMDLs are established to determine the pollutant load reductions needed to achieve and
maintain water quality standards. A water quality standard is the combination of a
designated use for a particular body of water and the water quality criteria designed to
protect that use. Designated uses include activities such as swimming, drinking water
supply, protection of aquatic life, and shellfish propagation and harvest. Water quality
criteria consist of narrative statements and numeric values designed to protect the
designated uses. Criteria may differ among waters with different designated uses.

The Upper Chester River watershed is associated with two assessment units in
Maryland's Integrated Report: a non-tidal 8-digit watershed (02130510) and an estuary
portion [Chesapeake Bay segment: Upper Chester River Tidal Fresh (CHSTF)]. This
TMDL only addresses the non-tidal portion of the watershed. Upstream loads from the
Delaware Chester River discharge into the non-tidal 8-digit watershed and are included in
this TMDL. A sediment TMDL for the CHSTF was established as part of the Chesapeake
Bay TMDLs in 2010. Background information on the tidal portion of the watershed is
presented for informational purposes only. Only a baseline load and load allocation have
been calculated for the Delaware Chester River

The Maryland Department of the Environment (MDE) identified the waters of the
Upper Chester River watershed and associated assessment units on the State's 2016
Integrated Report as impaired by multiple pollutants (MDE 2016). Table 1 identifies the
impairment listings associated with this watershed.

A data solicitation for sediment was conducted by MDE in March 2018 and all readily
available data have been considered.

Table 1: Upper Chester River Integrated Report Listings

Watershed	Basin Code	Tidal/Non-tidal	Designated Use Class	Year Listed	Identified Pollutant	Listing Category
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The Maryland Surface Water Use Designation in the Code of Maryland Regulations (COMAR) for the Upper Chester River watershed's non-tidal tributaries are designated as Use Class I - *water contact recreation and protection of aquatic life*. Tidal tributaries and the Upper Chester River mainstem are designated Use Class II - *support of estuarine and marine aquatic life and shellfish harvesting* (COMAR 2016a, b, c).

The Upper Chester River watershed was originally listed for biological impairment on the 2006 Integrated Report. The listing was based on the biological assessment methodology, which uses aquatic health scores, consisting of the Benthic Index of Biotic Integrity (BIBI) and Fish Index of Biotic Integrity (FIBI). These indices indicated that the biological metrics for the watershed exhibit a significant negative deviation from reference conditions (MDE 2006a).

In order to determine what stressor or stressors are impacting aquatic life, MDE's *Biological Stressor Identification* (BSID) methodology was applied. The BSID analysis for the Upper Chester River watershed identified sediment and instream habitat parameters. The sediment parameter shows a significant association with poor epifaunal substrate. The instream habitat parameter group shows a significant association with channelization present, marginal to poor instream habitat structure, poor riffle/run quality, poor velocity/depth diversity, and concrete/gabion present. Further details of this analysis are presented in the 2012 document entitled, *Watershed Report for Biological Impairment of the Non-Tidal Upper Chester River Watershed in Kent and Queen Anne's Counties, Maryland Biological Stressor Identification Analysis Results and Interpretation* (MDE 2013).

The objective of this TMDL is to ensure that watershed sediment loads are at a level that supports the Use Class I designation for the Upper Chester River watershed. The TMDL will address water clarity problems and associated impacts to aquatic life in the non-tidal Upper Chester River watershed caused by high sediment and TSS concentrations.

The CWA requires TMDLs to be protective of all the designated uses applicable to a particular waterbody. The primary focus of this TMDL is the designated use of protection of aquatic life because the Integrated Report listing was based on a biological assessment of the watershed. The biological assessment revealed the current levels of TSS and other pollutants prevent the watershed from achieving its designated use of supporting aquatic life. However, the required reductions are expected to protect all designated uses of the watershed, including water contact recreation. It is understood that aquatic life is more sensitive to sediment impacts than recreation because aquatic life impacts result from continuous exposure than can affect respiration and propagation. Recreation, on the other hand, is sporadic and sediment is unlikely to pose a human health risk due to dermal contact or minimal ingestion that would occur during recreation.

2.0 SETTING AND WATER QUALITY DESCRIPTION

2.1 General Setting

Location

The Upper Chester River and its tributaries are part of the Chester River basin that drains to the Chesapeake Bay. The Upper Chester River watershed covers 178 square miles in Maryland and Delaware. About 137 square miles of the Upper Chester River watershed are in Maryland, which includes 82 square miles in Queen Anne's County and 55 square miles in Kent County. Upper Chester River is approximately 16.9 kilometers (10.5 miles) in length and extends from the headwaters in Delaware downstream to the confluence with Foreman Branch, and the Middle Chester watershed extends from that point downstream to the confluence with Southeast Creek. The upper region of the Upper Chester River Watershed, near the Maryland and Delaware border, consists of uninhabited forests and wetlands, which are part of the Millington Wildlife Management Area. This is an area of approximately 3,800 acres, which drains into Cypress Branch, northeast of Millington. The Upper Chester River watershed is agriculturally diverse and high in the production of corn, wheat and soybean. The watershed area is located in the Coastal region of three distinct eco-regions identified in the Maryland Biological Stream Survey (MBSS) IBI metrics (Southerland et al. 2005) A location map of the Upper Chester River is provided in Figure 1.

According to the Chesapeake Bay Program's Phase 5.3.2 watershed model, the total drainage area of the Maryland 8-digit watershed is approximately 86,500 acres, not including water/wetlands. Approximately 300 acres of the watershed area is covered by water. The total population in the Upper Chester River watershed is approximately 7,200 (US Census Bureau 2010).

There are eight "high quality," or Tier II, stream segments located within the Upper Chester River watershed. Tier II segments are designated using MBSS data, and both the FIBI and BIBI values must be greater than 4.00 (on a scale of 1 – 5). Tier II segments require the implementation of Maryland's anti-degradation policy, which is designed to prevent degradation of high quality waters. The policy requires a review of all permitted activities upstream of Tier II stream segments. (COMAR 2016d; MDE 2011).

Geology/Soils

The Upper Chester River watershed is located solely within the Delmarva Peninsula Region of the Coastal Plain Province of Maryland. The Delmarva Peninsula Province encompasses the landmass between the Chesapeake Bay and the Delaware Bay. Wetlands are abundant in the Coastal Plain due to the low topographical relief and high groundwater characteristics of the region. The Coastal Plain province is characterized by unconsolidated sediments, which include sand, gravel, silt, and clay. The sediments of the coastal plain dip toward the east at a very low angle of 3 degrees, and some of the younger formations in the province crop out to the surface with increasing frequency in a southeasterly direction. The majority of the province, however, consists of older

formations, which are covered by a thin layer of Quaternary Gravel (MGS 2012). The two predominant soil types in the Upper Chester River watershed are the Sassafras and Ingleside soil associations. The Sassafras association makes up the majority of the northern portion of the Upper Chester River watershed, while the Ingleside association makes up the majority of the southern portion of the watershed. The Sassafras soil association is characterized by very deep, well drained, moderately permeable, fine-loam soils formed in sandy marine and old alluvial sediments of the Coastal Plain. The Ingleside soil association is characterized by very deep, well drained, moderately permeable, coarse-loam soils. Both Sassafras and Ingleside soils are categorized as prime farmland, which means that they are among the most productive soils in the state for agriculture and forestry. [U.S. Department of Agriculture (USDA) 2006].

The United States Department of Agriculture (USDA) Natural Resources Conservation Service (NRCS) classifies soils into 4 hydrologic soil groups: Group A soils have high infiltration rates and are typically deep well drained/excessively drained sands or gravels; Group B soils have moderate infiltration rates and consist of soils that are moderately deep to deep and moderately well to well drained soils, with moderately fine/coarse textures; Group C soils have slow infiltration rates with a layer that impedes downward water movement, and they primarily have moderately fine-to-fine textures; Group D soils have very slow infiltration rates consisting of clay soils with a permanently high water table that are often shallow over nearly impervious material. The Upper Chester River watershed is comprised primarily of Group B soils (54%) and Group C soils (30%), with small portions of the watershed consisting of Group A soils (4%) and Group D soils (12%) (USDA 2006).
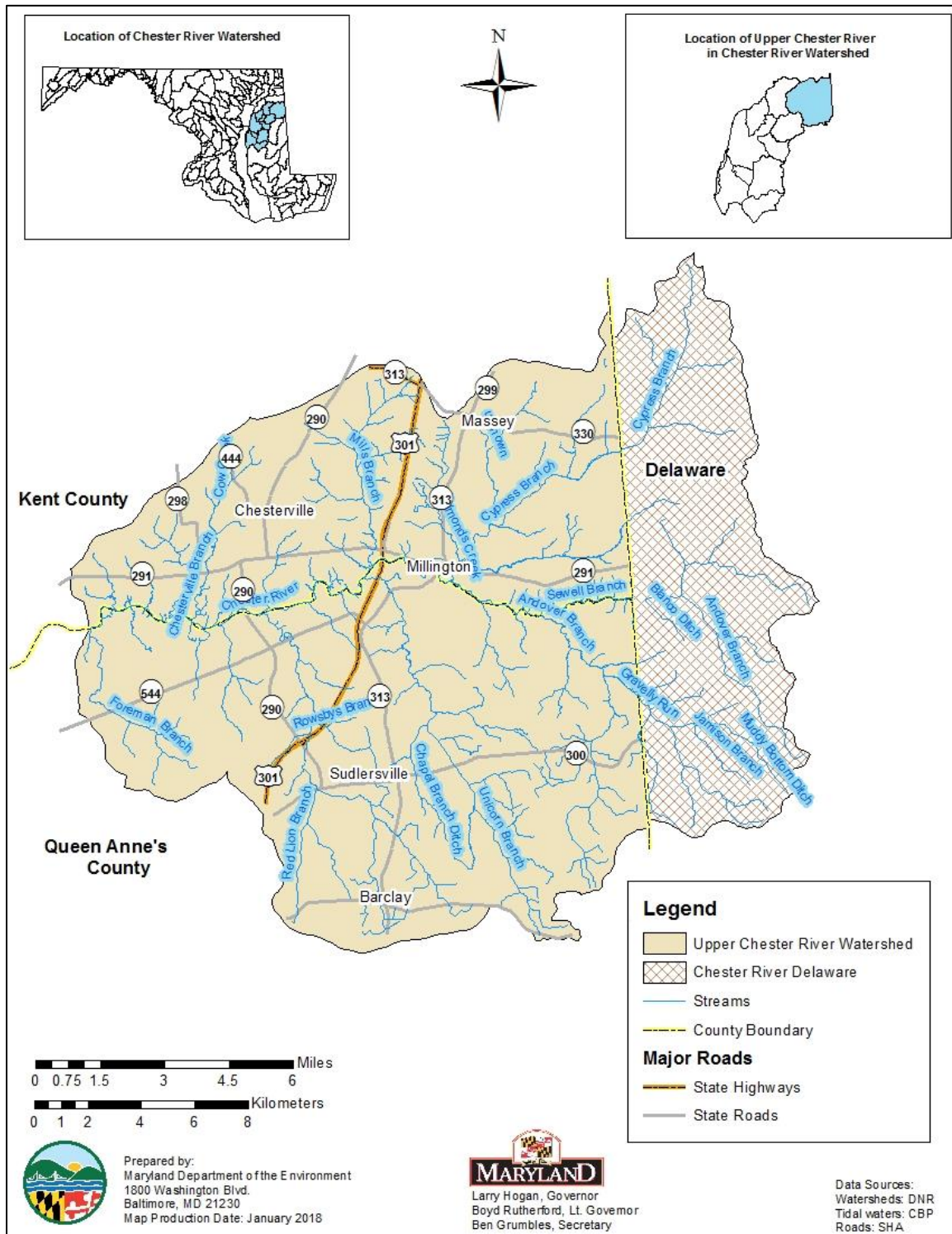
Figure 1: Location Map of the Upper Chester River Watershed in Kent and Queen Anne's Counties, Maryland

2.1.1 Land-use

Land-use Methodology

The land-use framework used to develop this TMDL was originally developed for the Chesapeake Bay Program Phase 5.3.2 (CBP P5.3.2) Watershed Model. The CBP P5.3.2 land-use was based on two distinct stages of development.

The first stage consisted of the development of the Chesapeake Bay Watershed Land-Cover Data (CBLCD) series of Geographic Information System (GIS) datasets. These datasets provide a 30-meter resolution raster representation of land-cover in the Chesapeake Bay watershed, based on sixteen Anderson Level two land-cover classes. The CBLCD basemap, representing 2001 conditions, was primarily derived from the Multi-Resolution Land Characteristics (MRLC) Consortium's National Land-Cover Data (NLCD) and the National Oceanic and Atmospheric Administration's (NOAA) Coastal Change Analysis Program's (CCAP) Land-Cover Data. By applying Cross Correlation Analysis to Landsat 5 Thematic Mapper and Landsat 7 Enhanced Thematic Mapper satellite imagery, CBLCD datasets for 1984, 1992, and 2006 from the baseline 2001 dataset. The watershed model documentation, *Chesapeake Bay Phase 5.3 Community Watershed Model* (USEPA 2010b), describes the development of the CBLCD series in more detail. USGS and NOAA also developed an impervious cover dataset from Landsat satellite imagery for the CBLCD basemap, which was used to estimate the percent impervious cover associated with CBLCD developed land-cover classifications.

The second stage consisted of using ancillary information for: 1) the creation of a modified 2006 CBLCD raster dataset, and 2) the subsequent development of the CBP P5.3.2 land-use framework in tabular format. Estimates of the urban footprint in the 2006 CBLCD were extensively modified using supplemental datasets. Navteq street data (secondary and primary roads) and institutional delineations were overlaid with the 2006 CBLCD land-cover and used to reclassify underlying pixels. Certain areas adjacent to the secondary road network were also reclassified based on assumptions developed by USGS researchers, in order to capture residential development (*i.e.*, subdivisions not being picked up by the satellite in the CBLCD). In addition to spatially modifying the 2006 CBLCD, the following datasets were used to supplement the developed land cover data in the final CBP P5.3.2 land-use framework: US Census housing unit data, Maryland Department of Planning (MDP) Property View data, and estimates of impervious coefficients for rural residential properties (determined via a sampling of these properties using aerial photography). This additional information was used to estimate the extent of impervious area in roadways and residential lots. Acres of construction and extractive land-uses were determined independently using a method developed by USGS. (Claggett, Irani, and Thompson 2012). Finally, in order to develop accurate agricultural land-use

acreages, the CBP P5.3.2 incorporated county level US Agricultural Census data (USDA 1982, 1987, 1992, 1997, 2002). The watershed model documentation, *Chesapeake Bay Phase 5.3 Community Watershed Model* (USEPA 2010b), describes these modifications in more detail.

The result of these modifications is that CBP P5.3.2 land-use does not exist in a single GIS coverage; instead, it is only available in a tabular format. The CBP P5.3.2 watershed model is comprised of 30 land-uses. The land-uses are divided into 13 classes with distinct sediment erosion rates. Table 2 lists the CBP P5.3.2 generalized land-uses, detailed land-uses, which are classified by their sediment erosion rates, and the acres of each land-use in the Upper Chester River watershed. The land-use acreage used to inform this TMDL is based on the CBP P5.3.2 2009 Progress Scenario.

Upper Chester River Watershed Land-Use Distribution

The land-use distribution of the Upper Chester River watershed consists primarily of agricultural (52%), forest (39%), and urban lands (5%). A detailed summary of the watershed land-use areas is presented in Table 2, and a land-use map is provided in Figure 2.

Table 2: Land-Use Percentage Distribution for the Upper Chester River Watershed

General Land Use	Detailed Land-Use	Area (Acres)	Percent (%)	Grouped Percent of Total (%)
	Forest	33,605	38.7	
Forest	Harvested Forest	338	0.4	39.1
Pasture	Pasture	2,771	3.2	3.2
Crop	Crop	44,791	51.6	51.6
AFO/CAFO	AFO/CAFO	20	0.0	0.0
Nursery	Nursery	348	0.4	0.4
Unregulated urban	Unregulated urban	4,598	5.3	5.3
	Construction	3	0.0	
	Developed	4	0.0	
Regulated Urban	Extractive	32	0.0	0.0
Water/wetlands	Water/wetlands	308	0.4	0.4
Total		**86,818**	**100.0**	**100.0**

1. Unregulated urban land use is not regulated by a NPDES permit. Regulated urban land use is regulated by an NPDES permit.
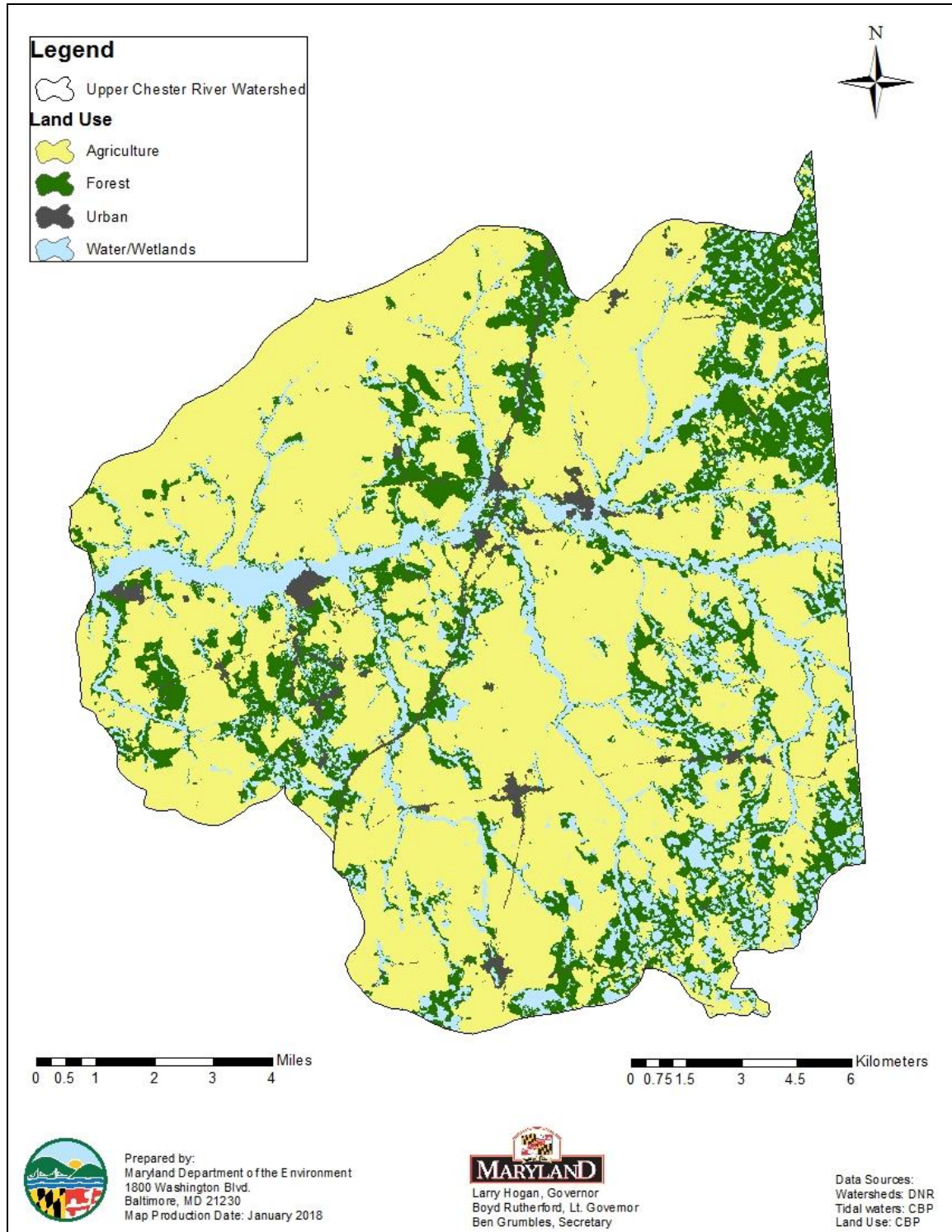2. Individual values may not add to total load due to rounding.

Figure 2: Land-use of the Upper Chester River Watershed

2.2 Source Assessment

The Upper Chester River Watershed Total Baseline Sediment Load consists of loads generated outside of the assessment unit, referred to as an Upstream Baseline Load, and loads generated within the assessment unit, referred to as the Upper Chester River Baseline Load Contribution. The Upper Chester River watershed baseline load contributions consists of nonpoint sources loads, and point source loads which can be further divided into National Pollutant Discharge Elimination System (NPDES) Stormwater loads, and Wastewater loads. This section summarizes the methods used to derive each of these distinct source categories.

2.2.1 Nonpoint Source Assessment

In this document, the nonpoint source loads account for all sediment loads not covered under a NPDES permit within the Upper Chester River watershed. In general, these are rainfall driven land-use based loads from agricultural and forested lands. In this watershed, unregulated urban land use is also included in the nonpoint source load. This section provides the background and methods for determining the nonpoint source baseline loads generated within the Upper Chester River watershed (Nonpoint Source BL_{UCR}).

General Load Estimation Methodology

Nonpoint source sediment loads generated within the Upper Chester River watershed are estimated based on the *edge-of-stream (EOS)* loads from the CBP P5.3.2 watershed model 2009 Progress Scenario. Within the CBP P5.3.2 watershed model, EOS sediment loads are calculated based on the fact that not all of the *edge-of-field* (EOF) sediment load is delivered to the stream or river (some of it is stored on fields down slope, at the foot of hillsides, or in smaller rivers or streams that are not represented in the model). To calculate the actual EOS loads, a *sediment delivery factor (SDF)* (the ratio of sediment reaching a basin outlet compared to the total erosion within the basin) is used. Details of the methods used to calculate sediment load have been documented in the report entitled *Chesapeake Bay Phase 5 Community Watershed Model* (USEPA 2010b). A summary of the methodology is presented in the following sections.

Edge-of-Field Target Erosion Rate Methodology

Edge-of-field erosion can be defined as erosion or sediment loss from any particular land surface. EOF target erosion rates are the values used in the calibration of the Chesapeake Bay Program (CBP) model, based on literature values. EOF target erosion rates for agricultural land-uses and forested land-use were based on erosion rates determined by the Natural Resource Inventory (NRI). The NRI is a statistical survey of land-use and natural resource conditions conducted by the Natural Resources Conservation Service (NRCS) (USDA 2006). The sampling methodology is explained by Nusser and Goebel (1997).

Estimates of average annual erosion rates for pasture and cropland are available on a county basis at five-year intervals, starting in 1982. Erosion rates for forested land-uses are not available on a county basis from the NRI; however, for the purpose of the Chesapeake Bay Program Phase 4.3 (CBP P4.3) watershed model, the NRI calculated average annual erosion rates for forested land-use on a watershed basis. These rates were used as targets in the CBP P5.3.2 model.

The average value of the 1982 and 1987 surveys was used as the basis for EOF target rates for pasture and cropland. Rates for urban pervious, urban impervious, extractive, and barren land were based on a combination of best professional judgment, literature analysis, and regression analysis. The EOF erosion rates do not reflect best management practices (BMPs) or other soil conservation policies introduced in the wake of the effort to restore the Chesapeake Bay. To compensate for this, BMPs are applied to the modeled EOS loads in the CBP P5.3.2 2009 Progress Scenario. BMP data, representing BMPs in place in 2009, was collected by the Chesapeake Bay Program (CBP), and TSS reduction efficiencies have been estimated by CBP for specific types of BMPs based on peer reviewed studies, data collected by local jurisdictions, and an analysis of available literature values. For further details regarding EOF erosion rates, please see Section 9.2.1 of the *Chesapeake Bay Phase 5 Community Watershed Model* (USEPA 2010b). Table 3 lists EOF erosion rates specific to Kent and Queen Anne's Counties, where the Upper Chester River watershed is located.

Table 3: Kent and Queen Anne's Counties Target EOF TSS Loading Rates (ton/acre/yr) by Land-Use

Land-use	Data Source	Kent County Target EOF TSS Loading rate (ton/acre/yr)	Queen Anne's County Target EOF TSS Loading rate (ton/acre/yr)
Forest	NRI (1987)	0.17	0.17
Harvested Forest	Literature values	3	3
Nursery	Equivalent to conventional till	5.58	4.33
Pasture	NRI average (1982-1987)	1.27	0.16
Animal Feeding Operations	NRI pasture average (1982-1987) multiplied by 9	11.43	1.44
Hay	Adjusted NRI average (1982-1987)	1.43	1.11
Conventional Till	Adjusted NRI average (1982 – 1987)	5.58	4.33
Conservation Till	Adjusted NRI average (1982 – 1987)	3.35	2.60
Pervious Urban	Regression Analysis	0.74	0.74
Extractive	Literature values/best professional judgment	10	10
Barren (Construction)	Literature values	23	23
Impervious Urban	Regression Analysis	5.18	5.18

Edge-of-Stream Sediment Loads

EOS sediment loads are the loads that enter the modeled river reaches. Modeled river reaches are those with discharges of 100 cubic feet per second (cfs) or greater. (Exceptions were made for some river reaches that had useful monitoring data but were less than 100 cfs.) EOS sediment loads represent not only the erosion from the land but all of the intervening processes of deposition on hillsides and sediment transport through smaller rivers and streams. The influence of the sum of these processes is represented in the estimated SDF.

The formula for the EOS load calculation within the CBP P5.3.2 watershed model is as follows:

$$\sum_{i}^{n} EOS = Acres_i * EOF_i * SDF_i$$
(Equation 2.1)

where:
n = number of land-use classifications
i = land-use classification
EOS = Edge of stream load, tons per year (ton/yr)
Acres = acreage for land-use i
EOF = Edge-of-field erosion rate for land-use i, ton/acre/yr
SDF = sediment delivery factor for land-use i

2.2.2 Point Source Assessment

A list of active permitted point sources that contribute to the sediment load in the Upper Chester River watershed was compiled using best available resources. The types of permits identified were individual municipal permits, general mining permits, general stormwater permits, and the general permit for stormwater discharges from construction sites. The permits can be grouped into two categories: wastewater and stormwater. The wastewater category includes those loads generated by continuous discharge sources whose permits have TSS limits. Wastewater permits that do not meet these conditions are considered *de minimis* in terms of the total sediment load. The stormwater category includes all NPDES regulated stormwater discharges. The technical memorandum to this document, entitled *Point Sources of Sediment in the Non-Tidal Upper Chester River Watershed,* identifies all the wastewater permits and NPDES regulated stormwater discharges that contribute to the sediment load in the Upper Chester River watershed. There are no wastewater or stormwater permits in the upstream Delaware Chester River.

The baseline sediment loads for the wastewater permits (Wastewater BL_{UCR}) are calculated based on their permitted TSS limits (average monthly or weekly concentration values) and corresponding flow information. The stormwater permits identified throughout the Upper Chester River watershed do not include numeric TSS limits. In the absence of TSS limits, the NPDES regulated stormwater baseline load (NPDES

Stormwater BL_{UCR}) is calculated using the CBP P5.3.2 Progress Scenario urban land-use EOS loads (as per Equation 2.1) similar to the approach for NPS loads outlined in Section 2.1. The technical memorandum to this document entitled *Point Sources of Sediment in the Non-Tidal Upper Chester River Watershed* provides detailed information regarding the calculation of the Upper Chester River watershed Wastewater BL_{UCR} and NPDES Stormwater BL_{UCR}.

2.2.3 Upstream Loads Assessment

For the purpose of this analysis, one upstream watershed has been identified: the Delaware Chester River watershed. Subsequently, the sediment baseline loads from this upstream watershed will be presented as a Delaware Chester River Baseline Load (BL_{DE}). The BL_{DE} is estimated based on the same nonpoint source load estimation methodology described in Section 2.2.1.

2.2.4 Summary of Baseline Loads

Table 4 summarizes the Upper Chester River Baseline Sediment Load, reported in tons per year (ton/yr) and presented in terms of Upstream Baseline Loads and Upper Chester River Baseline Load Contributions: Nonpoint source, NPDES Stormwater, and Wastewater Baseline Loads

Table 4: Upper Chester River Watershed Baseline Sediment Loads (ton/yr)

		Upstream Baseline Load		Patuxent River Lower Watershed Baseline Load Contribution				
Total Baseline Load	=	**BL_{DE}**	+	**Nonpoint Source BL_{UCR}**	+	**NPDES Stormwater BL_{UCR}**	+	**Process Water BL_{UCR}**
7,150	=	1,170	+	5,955	+	15	+	10

Table 5 presents a breakdown of Upper Chester River Watershed Total Baseline Sediment Load, detailing loads per land-use or other source category. Upstream loads are not included in Table 5.

Table 5: Detailed Baseline Sediment Loads Within the Upper Chester River Watershed

General Land Use	Detailed Land-Use	ton/yr	Percent (%)	Grouped Percent of Total (%)
Forest	Forest	311	5.2	
	Harvested Forest	22	0.4	5.6
AFO/CAFO	Animal Feeding Operations	4	0.1	0.1
Pasture	Pasture	70	1.2	1.2
Crop	Crop	5,081	85.0	85.0
Nursery	Nursery	139	2.3	2.3
Unregulated Urban	Unregulated Urban	329	5.5	5.5
Regulated Urban	Construction	2	0.0	
	Developed	1	0.0	
	Extractive	12	0.2	0.2
Point Sources	Industrial Point Sources	0	0.0	
	Municipal Point Sources	10	0.2	0.2
Total		**5,980**	**100.0**	**100.0**

Note: Individual values may not add to total load due to rounding.

2.3 Water Quality Characterization

The non-tidal Upper Chester River watershed was originally listed for impacts to biological communities in the 2006 Integrated Report. To refine the listing for impacts to biological communities, Maryland conducted a stressor identification analysis. Details of this analysis are presented below and in the document entitled, *Watershed Report for Biological Impairment of the Upper Chester River Watershed in Kent and Queen Anne's Counties, Maryland Biological Stressor Identification Analysis Results and Interpretation* (MDE 2012).

Currently in Maryland, there are no specific numeric criteria for suspended sediments. Therefore, to determine whether aquatic life is impacted by elevated sediment loads, MDE's BSID methodology was applied. The primary goal of the BSID analysis is to identify the most probable cause(s) for observed biological impairments throughout MD's 8-digit watersheds (MDE 2009a).

The BSID analysis applies a case-control, risk-based, weight-of-evidence approach to identify potential causes of biological impairment. The risk-based approach estimates the strength of association between various stressors and an impaired biological community. The BSID analysis then identifies individual stressors as probable or unlikely causes of

the poor biological conditions within a given watershed, and subsequently reviews ecological plausibility. Finally, the analysis concludes whether or not these individual stressors or groups of stressors are contributing to the impairment (MDE 2009a).

The primary dataset for BSID analysis includes Maryland Department of Natural Resources (MDDNR)-MBSS Round 2 and Round 3 data (collected between 2000-2009) because it provides a broad spectrum of paired data variables, which allow for a more comprehensive stressor analysis. MDDNR-MBSS Round 1 can also be used if there is limited Round 2 and 3 data. The MBSS is a robust statewide probability-based sampling survey for assessing the biological conditions of 1st through 4th order, non-tidal streams (Klauda et al. 1998; Roth et al. 2005). It uses a fixed length (75 meter) randomly selected stream segment for collecting site level information within a primary sampling unit (PSU), also defined as a watershed. The randomly selected stream segments, from which field data are collected, are selected using either stratified random sampling with proportional allocation, or simple random sampling (Cochran 1977). The random sample design allows for unbiased estimates of overall watershed conditions. Thus, the dataset facilitated case-control analyses because: 1) in-stream biological data are paired with chemical, physical, and land-use data variables that could be identified as possible stressors; and 2) it uses a probabilistic statewide monitoring design.

The BSID analysis combines the individual stressors (physical and chemical variables) into three generalized parameter groups in order to assess how the resulting impacts of these stressors can alter the biological community and structure. The three generalized parameter groups include: sediment, habitat, and water chemistry. Identification of a sediment stressor as contributing to the biological impairment is based on the results of the individual stressor associations within the sediment parameter grouping, which reveal the effects of sediment related impacts on stream biota (MDE 2009a).

Upper Chester River Watershed Monitoring Stations

A total of 30 water quality monitoring stations were used to characterize the Upper Chester River watershed for the purpose of this TMDL. The biological assessment was based on the combined results of MBSS Round 1 and Round 2 data, which includes 30 stations. The BSID analysis used stations from MBSS Round 2, which includes 10 stations. All stations are listed in Table 6 and presented in Figure 3.

Table 6: Monitoring Stations in the Upper Chester River Watershed

| Site Number	Sponsor	Site Type	Location	Latitude (decimal degrees)	Longitude (decimal degrees)
KE-N-046-226-95	DNR	MBSS Round 1	Sewell Branch	39.2444	-75.7831
KE-N-116-123-95	DNR	MBSS Round 1	Andover Branch, unnamed tributary	39.2554	-75.8274
QA-N-007-217-95	DNR	MBSS Round 1	Unicorn Branch	39.2257	-75.8536
QA-N-024-209-95	DNR	MBSS Round 1	Gravelly Run	39.2150	-75.7571
QA-N-033-301-95	DNR	MBSS Round 1	Red Lion Branch	39.2251	-75.9035
QA-N-033-304-95	DNR	MBSS Round 1	Red Lion Branch	39.1948	-75.8993
QA-N-033-307-95	DNR	MBSS Round 1	Red Lion Branch	39.1999	-75.8976
QA-N-033-309-95	DNR	MBSS Round 1	Red Lion Branch	39.2299	-75.9038
QA-N-033-310-95	DNR	MBSS Round 1	Red Lion Branch	39.2312	-75.9040
QA-N-033-311-95	DNR	MBSS Round 1	Red Lion Branch	39.1916	-75.9007
QA-N-033-313-95	DNR	MBSS Round 1	Red Lion Branch	39.1891	-75.8992
QA-N-033-314-95	DNR	MBSS Round 1	Red Lion Branch	39.2181	-75.9006
QA-N-033-317-95	DNR	MBSS Round 1	Red Lion Branch	39.1954	-75.8991
QA-N-033-318-95	DNR	MBSS Round 1	Red Lion Branch	39.1935	-75.8999
QA-N-033-321-95	DNR	MBSS Round 1	Red Lion Branch	39.2324	-75.9047
QA-N-041-109-95	DNR	MBSS Round 1	Andover Branch	39.1945	-75.7752
QA-N-041-113-95	DNR	MBSS Round 1	Andover Branch	39.1670	-75.7849
QA-N-059-125-95	DNR	MBSS Round 1	Unicorn Branch, unnamed tributary	39.2039	-75.8213
QA-N-111-312-95	DNR	MBSS Round 1	Andover Branch	39.2284	-75.7798
QA-N-111-315-95	DNR	MBSS Round 1	Andover Branch	39.2272	-75.7789
UPCR-101-R-2004	DNR	MBSS Round 2	Mills Branch	39.3002	-75.8809
UPCR-102-R-2004	DNR	MBSS Round 2	Chester River, unnamed tributary 2	39.2759	-75.9389
UPCR-104-R-2004	DNR	MBSS Round 2	Chester River, unnamed tributary 1	39.2550	-75.9689
UPCR-106-R-2004	DNR	MBSS Round 2	Andover Branch, unnamed tributary	39.2400	-75.8188
UPCR-107-R-2004	DNR	MBSS Round 2	Red Lion Branch, unnamed tributary 2/3	39.1608	-75.8532
UPCR-112-R-2004	DNR	MBSS Round 2	Red Lion Branch, unnamed tributary 2/3	39.1673	-75.8598
UPCR-115-R-2004	DNR	MBSS Round 2	Red Lion Branch, unnamed tributary 1	39.2050	-75.8918
UPCR-116-R-2004	DNR	MBSS Round 2	Red Lion Branch, unnamed tributary 2/3	39.1657	-75.8585
UPCR-208-S-2000	DNR	MBSS Round 2	Cypress Branch	39.2877	-75.7836
UPCR-214-R-2004	DNR	MBSS Round 2	Red Lion Branch, unnamed tributary 2	39.1812	-75.8842
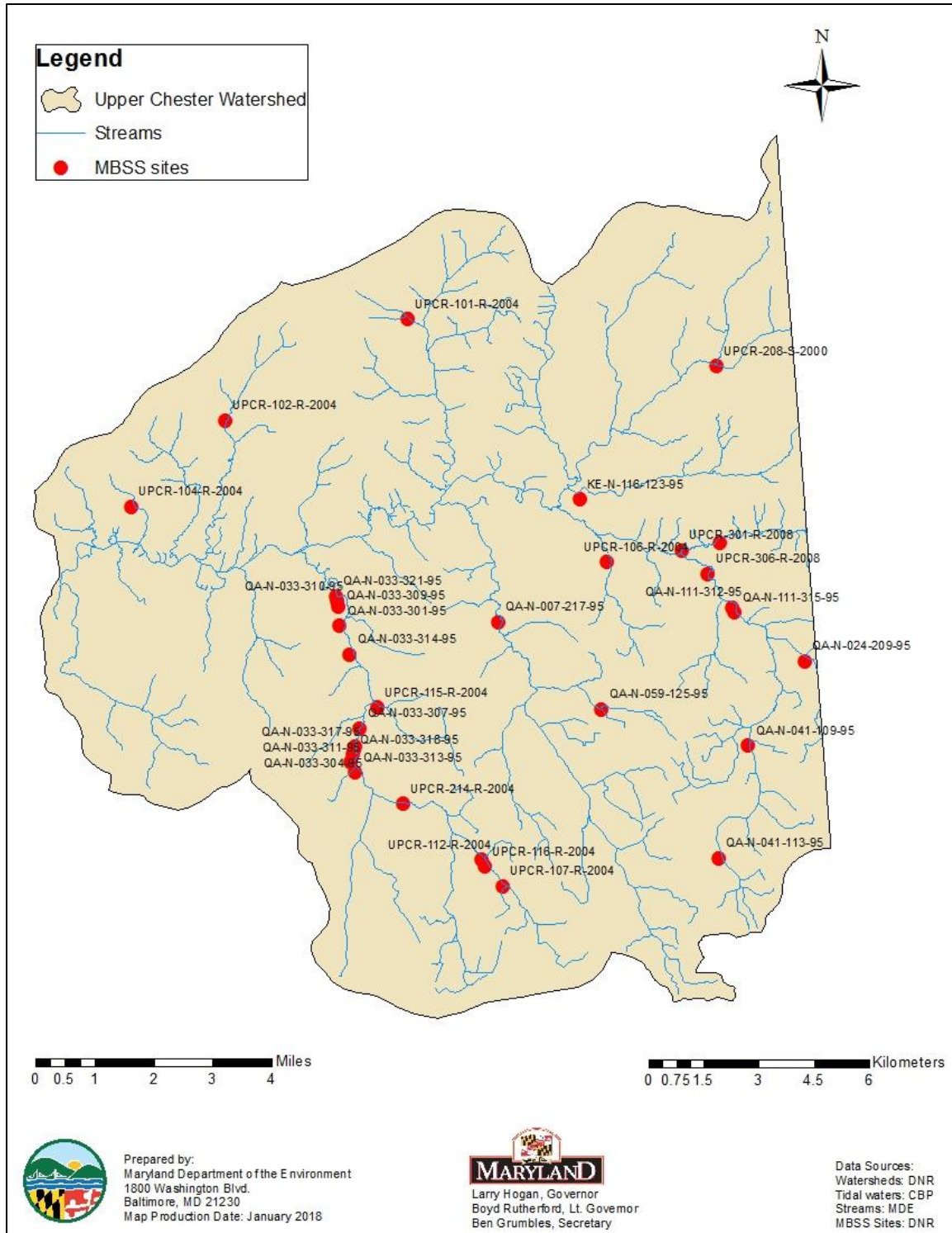
Figure 3: Monitoring Stations in the Upper Chester River Watershed

2.4 Water Quality Impairment

The Maryland Surface Water Use Designation in the COMAR for the Upper Chester River watershed's non-tidal streams are Use Class I - *water contact recreation, and protection of non-tidal warmwater aquatic life.* All of the tidal waters are designated Use Class II - *support of estuarine and marine aquatic life and shellfish harvesting* (COMAR 2016a, b, c). A map of the Designated Use Classes is provided in Figure 4.

This TMDL only addresses the non-tidal portion of the watershed. Upstream loads from the Delaware Chester River discharge into the non-tidal 8-digit watershed and are included in this TMDL. A sediment TMDL for the tidal portion of the watershed, CHSTF, was established as part the Chesapeake Bay TMDLs in 2010.

The water quality impairment of the Upper Chester River watershed addressed by this TMDL is caused, in part, by an elevated sediment load beyond a level that the watershed can sustain; thereby causing sediment related impacts to aquatic life. Assessment of aquatic life is based on BIBI and FIBI scores, as demonstrated via the BSID analysis for the watershed.

The Upper Chester River watershed was originally listed on Maryland's 2006 Integrated Report as impaired for impacts to biological communities. The biological assessment was based on the combined results of MBSS Round 1 (1995-1997) and Round 2 (2000-2004) data, which included 30 stations. 32% of the stream miles in the watershed, were assessed as having BIBI and/or FIBI scores significantly lower than 3.0 (on a scale of 1 to 5) (MDE 2006). See Figure 3 and Table 6 for station locations and information.

The results of the BSID analysis for the Upper Chester River watershed are presented in a report entitled *Watershed Report for Biological Impairment of the Non-Tidal Upper Chester River Watershed in Kent and Queen Anne's Counties, Maryland Biological Stressor Identification Analysis Results and Interpretation*. The report states that the degradation of biological communities in the Upper Chester River watershed is strongly associated with agricultural impacts, poor epifaunal substrate, channelization, and high total nitrogen (MDE 2012).

The BSID analysis determined that the biological impairment in the Upper Chester River watershed is due in part to stressors within the sediment parameter grouping. Overall, stressors within the sediment parameter grouping were identified as having a statistically significant association with impaired biological communities at approximately 33% of the sites with BIBI and/or FIBI scores significantly less than 3.0 throughout the watershed (MDE 2013). Therefore, since sediment is identified as a stressor to the biological communities in the Upper Chester River watershed, the watershed was listed as impaired by sediment in the 2012 Integrated Report, and a TMDL is required.
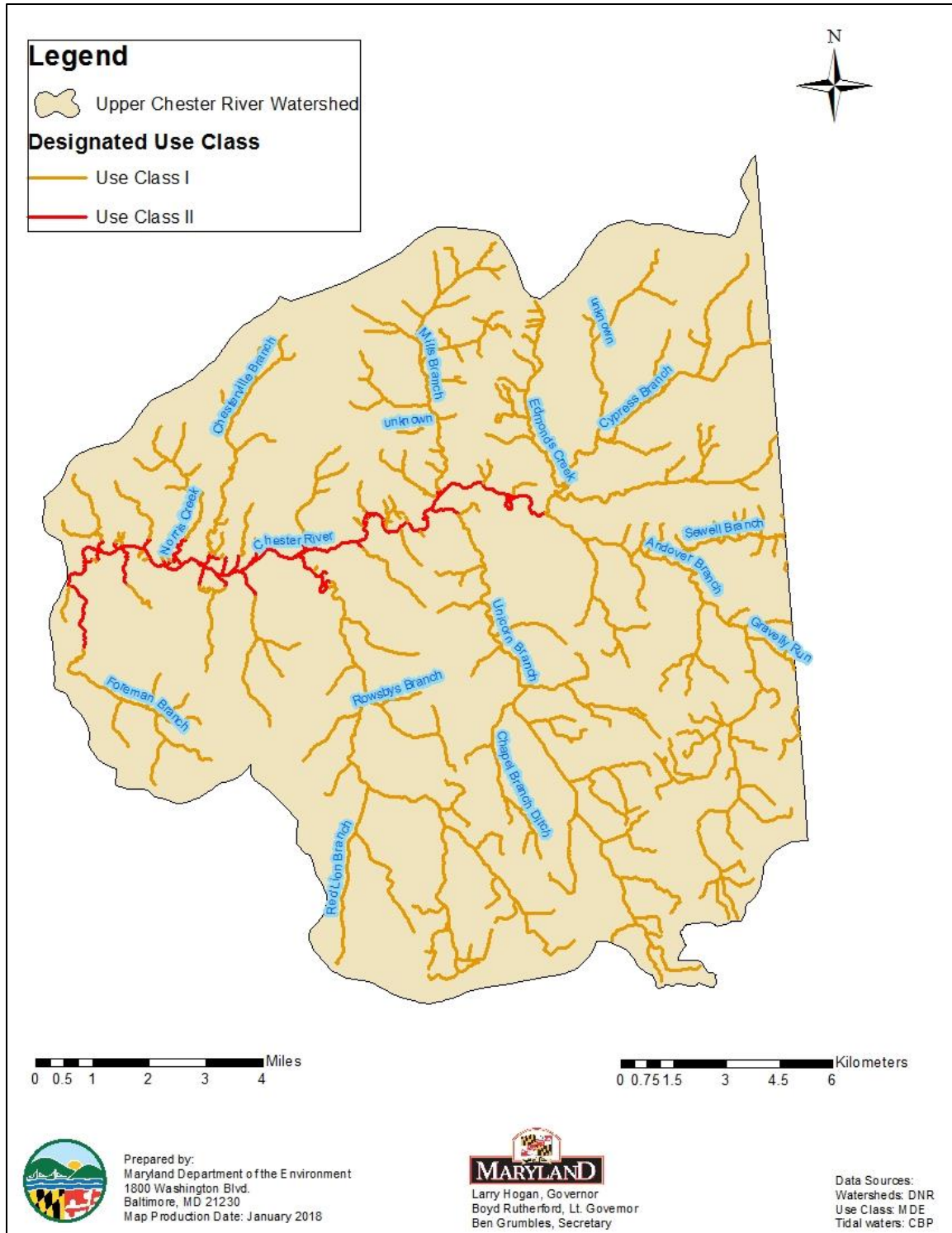
Figure 4: Designated Use Classes of the Upper Chester River Watershed in Kent and Queen Anne's Counties, Maryland

3.0 TARGETED WATER QUALITY GOAL

The objective of the sediment TMDL established herein is to reduce sediment loads, and their detrimental effects on aquatic life in the Upper Chester River watershed, to levels that support the Use Class I designation for the watershed. Excessive sediment has been identified by the USEPA as the leading cause of impairment of our nation's waters, and as contributing to the decline of populations of aquatic life in North America (USEPA 2003a). Sediment in streams may reduce visibility, preventing fish from seeing their prey, and may clog gills and filter feeding mechanisms of fish and benthic (bottom-dwelling) organisms. Excessive deposition of sediment on streambeds may bury eggs or larvae of fish and benthic macroinvertebrates, or degrade habitat by clogging the interstitial spaces between sand and gravel particles. Excessive sediment can also create hazards for recreation due to low visibility and the possibility of unseen objects.

Reductions in sediment loads are expected to result from decreased watershed erosion, which will then lead to improved benthic and fish habitat conditions. Specifically, sediment load reductions are expected to result in an increase in the number of benthic sensitive species present, an increase in the available and suitable habitat for a benthic community, a decrease in fine sediment (fines), and improved stream habitat diversity, all of which will result in improved water quality.

The TMDL will not completely resolve the impairment to biological communities within the watershed, since the BSID watershed analysis also identifies additional possible stressors impacting the biological conditions (e.g. channelization). This impairment to aquatic life will only be fully addressed when all substances identified as impairing biological communities in the watershed are reduced to levels that will meet water quality standards. (MDE 2009a, 2013).

4.0 TOTAL MAXIMUM DAILY LOADS AND SOURCE ALLOCATION

4.1 Overview

This section describes how the sediment TMDL and the corresponding allocations were developed for the Upper Chester River watershed.

4.2 Analysis Framework

Since there are no specific numeric criteria in Maryland that quantify the impact of sediment on the aquatic life of non-tidal stream systems, a reference watershed approach was used to establish the TMDL. In order to use a reference watershed approach, sediment loads were estimated using a watershed model. For this analysis, the CBP P5.3.2 model was used to calculate the sediment loads used in the reference watershed approach.

Watershed Model

The CBP P5.3.2 watershed model was chosen to estimate the sediment loads for the Upper Chester River watershed TMDL and the loads were expressed as EOS sediment loads. The spatial domain of the CBP P5.3.2 watershed model segmentation aggregates to the MD 8-digit watersheds, which is with the scale of the impairment listing. The nonpoint source baseline sediment loads generated within the Upper Chester River watershed are based on the EOS loads from the CBP P5.3.2 watershed model 2009 Progress Scenario. CBP P5.3.2 Progress Scenario EOS loads are calculated as the sum of individual land-use EOS loads within the watershed and represent a long-term average loading rate. Individual land-use EOS loads are calculated within the CBP P5.3.2 watershed model as a product of the land-use area, land-use target EOF loading rate, and loss from the EOF to the main channel. BMP data and reduction efficiencies are then subsequently applied to produce the final EOS loads. The loss from the EOF to the main channel is the *sediment delivery factor* and is defined as the ratio of the sediment load reaching a basin outlet to the total erosion within the basin. A *sediment delivery factor* is estimated for each land-use type based on the proximity of the land-use to the main channel. Thus, as the distance to the main channel increases, more sediment is stored within the watershed (i.e., *sediment delivery factor* decreases). Details of the data sources for the unit loading rates can be found in Section 2.2 of this report.

Reference Watershed Approach

In order to quantify the impact of sediment on the aquatic life of non-tidal stream systems, a reference watershed TMDL approach was used. Reference watersheds are those watersheds that are identified as supporting aquatic life, based on Maryland's biological assessment methodology. The biological assessment methodology assesses biological impairment at the watershed scale based on the percentage of MBSS monitoring stations, translated into watershed stream miles, that have BIBI and/or FIBI scores lower than the Minimum Allowable IBI Limit (MAL). The MAL represents the threshold under which a watershed is listed as impaired for biology and is calculated

based on the average annual allowable IBI value of 3.0 (on a scale of 1 to 5), the coefficient of variation of annual sentinel site results, and an assumed normal distribution. It accounts for annual variability and helps to avoid classification errors (i.e., false positives) when assessing for biological impairments (Roth et al. 1998, 2000; Stribling et al. 1998; MDE 2014b). For a full description of the selection of reference watersheds, please see *A Methodology for Addressing Sediment Impairments in Maryland's Non-tidal Watersheds* (MDE 2006).

Comparison of sediment loads from impaired watersheds to loads from reference watersheds requires that the watersheds be similar in physical and hydrological characteristics. For the establishment of this specific TMDL, watersheds were selected from the non-tidal Coastal Plain region since the Upper Chester River watershed is within this geologic province (see Section 2.1). See Appendix A for the list of reference watersheds. The same methodology as described in MDE 2006 for the selection of the Highland and Piedmont reference watersheds was used to select the Coastal Plain reference watersheds. Furthermore, all subsequent methodologies used to establish the TMDL end point, based on these reference watersheds, are exactly the same as those described in MDE 2006.

To further reduce the effect of the variability within the Coastal Plain physiographic regions (i.e., soils, slope, etc.), the watershed sediment loads were then normalized by a constant background condition, the all forested watershed condition. This new normalized term, defined as the *forest normalized sediment load* (Y_n), represents how many times greater the current watershed sediment load is than the *all forested sediment load* (y_{for}). The y_{for} is a modeled simulation of what the sediment load would be if the watershed were in its natural all forested state, instead of its current mixed land use. It is calculated using the CBP P5.3.2 model. The *forest normalized sediment load* for this TMDL is calculated as the baseline watershed sediment load divided by the *all forested sediment load*. The equation for the *forest normalized sediment load* is as follows:

$$Y_n = \frac{y_{ws}}{y_{for}}$$ (Equation 4.1)

Where:
  Y_n = forest normalized sediment load
  y_{ws} = current watershed sediment load (ton/yr)
  y_{for} = all forested sediment load (ton/yr)

FINAL

Six reference watersheds were identified in the Coastal Plain physiographic region. Reference watershed *forest normalized sediment loads* were calculated using CBP P5.3.2 watershed model 2009 Progress Scenario EOS loads. The median and 75th percentile of the reference watershed *forest normalized sediment loads* were calculated and found to be 5.9 and 8.8 respectively[1]. The median value of 5.9 was used as an environmentally conservative approach for establishing the sediment loading threshold for the TMDL (see Appendix A for more details).

The *forest normalized sediment load* for the Upper Chester River watershed, estimated as 7.4, was calculated using CBP P5.3.2 2009 Progress Scenario EOS loads, as follows:

$$Y_n = \frac{y_{ws}}{y_{for}} = \frac{5,970 \; ton/yr}{860 \; ton/yr} = 7.4 \qquad \text{(Calculation 4.1)}$$

A comparison of the Upper Chester River watershed *forest normalized sediment loads* to the *sediment loading threshold* demonstrates that the watershed exceeds the *sediment loading threshold*, indicating that it is receiving loads above the maximum allowable load that it can sustain and still meet water quality standards.

4.3 Scenario Descriptions and Results

The following analyses compare baseline conditions in the watershed (under which water quality problems exist) with potential future conditions, which project the water quality response to various simulated sediment load reductions. The analyses are grouped according to baseline conditions and future conditions associated with TMDLs.

<u>Baseline Conditions</u>

The baseline conditions are intended to provide a point of reference by which to compare the future scenario that simulates conditions of a TMDL. Baseline loads are calculated for nonpoint and point source loads. Point source loads can be subdivided into two categories, wastewater and stormwater.

The Upper Chester River watershed baseline nonpoint source sediment loads are estimated using the land-use and EOS sediment loading rates from the CBP P5.3.2 2009 Progress Scenario. The 2009 Progress Scenario was chosen because it is used as the baseline year in the Chesapeake Bay TMDL. The 2009 Progress Scenario represents 2009 land-use and BMP implementation simulated using precipitation and other meteorological inputs from the period 1990-2000 to represent variable hydrological

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[1] The 75th percentile value of reference condition streams was recommended by EPA to be used in establishing numerical criteria (MDE 2006). The median was found, for the sediment reference watersheds, to be approximately equivalent to other more complex statistical analyses and was used for ease of calculation (MDE 2009b). Both of these values ensure that the selected threshold will represent the reference group values, with the median being more conservative (lower).

conditions, thereby addressing annual changes in hydrology and capturing wet, average and dry years. The period 1991-2000 is the hydrological simulation period for the Chesapeake Bay TMDL.

The wastewater point source baseline sediment loads are estimated based on the existing permit information. The stormwater point source baseline sediment loads are also based on CBP 5.3.2 loading rates, specifically those for urban land use. Details of these loading source estimates can be found in Section 2.2 and the technical memorandum to this document entitled *Point Sources of Sediment in the Upper Chester River Watershed*.

TMDL Conditions

The TMDL scenario simulates conditions under which sediment loads have been reduced to levels that support aquatic life. In the TMDL calculation, the allowable load for the impaired watershed is calculated as the product of the *sediment loading threshold* (determined from watersheds with a healthy biological community) and the Upper Chester River watershed *all forested sediment load* (see Section 4.2). The resulting load is considered the maximum allowable load the watershed can sustain and support aquatic life.

The TMDL loading and associated reductions are averaged at the watershed scale; however, it is important to recognize that some subwatersheds may require higher reductions than others, depending on the distribution of the land-use.

The formula for estimating the TMDL is as follows:

$$TMDL = \sum_{i=1}^{n} Yn_{ref} \cdot y_{for_i} \qquad \text{(Equation 4.2)}$$

Where:
 TMDL = allowable load for impaired watershed (ton/yr)
 Yn_{ref} = sediment loading threshold
 y_{for_i} = all forested sediment load for CBP P5.3.2 model segment i (ton/yr)
 i = CBP P5.3.2 model segment
 n = number of CBP P5.3.2 model segments in watershed

4.4 Critical Condition and Seasonality

USEPA's regulations require TMDLs to take into account seasonality and critical conditions for stream flow, loading, and water quality parameters (CFR 2012b). The intent of this requirement is to ensure that the water quality of the waterbody is protected during times when it is most vulnerable. The biological monitoring data used to determine the reference watersheds reflect the impacts of stressors (i.e., sediment impacts to stream biota) over the course of time and therefore depict an average stream condition

(i.e., captures all high and low flow events). Since the TMDL endpoint is based on the median of forest normalized loads from watersheds assessed as having good biological conditions (i.e., passing Maryland's biological assessment), by the nature of the biological data described above, it must inherently include the critical conditions of the reference watersheds. Therefore, since the TMDL reduces the watershed sediment load to a level compatible with that of the reference watersheds, critical conditions are inherently addressed. Moreover, the sediment loading rates used in the TMDL were determined using the CBP P5.3.2 model, which is a continuous simulation model with a simulation period 1991-2000, based on Hydrological Simulation Program Fortran (HSPF) model, thereby addressing annual changes in hydrology and capturing wet, average, and dry years.

Seasonality is captured in two components. First, it is implicitly included through the use of the biological monitoring data as this data reflects the impacts of stressors over time, as described above. Second, the MBSS dataset included benthic sampling in the spring (March 1 - April 30) and fish sampling in the summer (June 1 - September 30). Benthic sampling in the spring allows for the most accurate assessment of the benthic population, and therefore provides an excellent means of assessing the anthropogenic effects of sediment impacts on the benthic community. Fish sampling is conducted in the summer when low flow conditions significantly limit the physical habitat of the fish community, and it is therefore most reflective of the effects of anthropogenic stressors as well.

4.5 TMDL Loading Caps

This section presents the Upper Chester River watershed average annual sediment TMDL. This load is considered the maximum allowable long-term average annual load the watershed can sustain and support aquatic life.

The long-term average annual TMDL was calculated for the Upper Chester River watershed based on Equation 4.2 and set at a load 5.9 times the all forested condition of the watershed. In order to attain the TMDL loading cap calculated for the watershed, reductions were applied to the predominant sediment sources (i.e., significant contributors of sediment to the stream system), independent of jurisdiction. Sediment reductions are also required in the Upper Chester River watershed to meet the sediment allocations assigned under the 2010 Chesapeake Bay TMDL for sediment in the Upper Chester River Tidal Fresh Water Quality Segment. To ensure consistency with the Bay TMDL, and therefore efficiency in the reduction of sediment loads, reductions will be applied to the same sediment sources identified in Maryland's Watershed Implementation Plans (WIPs) for the Bay TMDL, as applicable in the watershed. These include: (1) regulated developed land; (2) unregulated urban land, (3) conventional till crops, conservation till crops, hay, and pasture; (4) harvested forest; (5) unregulated animal feeding operations and concentrated animal feeding operations (CAFOs); and (6) industrial wastewater sources and municipal wastewater treatment plants. Forest land is not assigned reductions because it is considered the most natural condition in the watershed.

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The Upper Chester River Watershed Baseline Load and TMDL are presented in Table 7.

Table 7: Upper Chester River Watershed Baseline Load and TMDL

Baseline Load (ton/yr)	TMDL (ton/yr)	Total Reduction (%)
7,150	5,673	21

Note: The load summary includes the Delaware Chester River Upstream Baseline
Load and TMDL Load Allocation.

4.6 Load Allocations Between Nonpoint and Point Sources

Per USEPA regulation, all TMDLs need to be presented as a sum of Wasteload
Allocations (WLAs) for point sources and Load Allocations (LAs) for nonpoint source
loads generated within the assessment unit, accounting for natural background, tributary,
and adjacent segment loads (CFR 2012a). The State reserves the right to allocate the
TMDL among different sources in any manner that protects aquatic life from sediment
related impacts.

<u>Load Allocation</u>
Individual LAs for each nonpoint land-use sector were calculated using the allocation
methodology in the MD Phase I Watershed Implementation Plan (WIP), which was
designed to be equitable, effective, and consistent with water quality standards (MDE
2010). The allocations were calculated by applying equal reductions to the *reducible*
loads of all sectors. The *reducible* load is defined as the difference between the No
Action (NA) scenario and the "Everything, Everyone, Everywhere" (E3) scenario. The
NA scenario represents current land-uses without any sediment controls applied, while
the E3 scenario represents the application of all possible BMPs and control technologies
to current land-use. For more detailed information regarding the calculation of the LA,
please see *Maryland's Phase I Watershed Implementation Plan for the Chesapeake Bay
Total Maximum Daily Load.*

In this watershed, unregulated urban land, crop land, nursery, and pasture were identified
as the predominant nonpoint sources of sediment and require reductions. Other land uses
that individually contributed less than 1% of the total sediment load were not reduced as
they would produce no discernible reductions. Forest is not assigned reductions, as it
represents the most natural condition in the watershed. Sediment loads from regulated
urban lands under National Pollutant Discharge Elimination System (NPDES) permits are
considered point source loads that must be included in the WLA portion of a TMDL
(USEPA 2002).

In this document, the LA for the Upper Chester River watershed is expressed as one
aggregate value for all nonpoint sources, including unregulated urban land use. For more
detailed information regarding the Upper Chester River watershed TMDL nonpoint
source LA, please see the technical memorandum to this document entitled *Nonpoint
Sources of Sediment in the Upper Chester River Watershed.*

A summary of the baseline and load allocation for nonpoint sources is presented in Table 8. The percent reduction shown in Table 8 does not represent the reduction applied to reducible loads, but the required reduction between the allocation and the baseline load.

Wasteload Allocation

The WLA of the Upper Chester River watershed is allocated to two permitted source categories, the Wastewater WLA and the Stormwater WLA. The categories are described below.

Wastewater WLA

Wastewater permits with specific TSS limits and corresponding flow information are assigned a WLA. In this case, detailed information is available to accurately estimate the WLA. If specific TSS limits are not explicitly stated in the wastewater permit, then TSS loads are expected to be *de minimis*. If loads are *de minimis,* they pose little risk to the aquatic environment.

Wastewater permits with specific TSS limits include:

- Individual industrial facilities

- Individual municipal facilities

- General mineral mining facilities

There is one minor municipal wastewater source with explicit TSS limits in the Upper Chester River watershed that contributes to the watershed sediment load. The total estimated TSS load from the wastewater source is based on current, average permit limits and is equal to 10 ton/yr. For more detailed information on the wastewater permits, please see the technical memorandum entitled *Point Sources of Sediment in the Upper Chester River Watershed*.

Stormwater WLA

NPDES stormwater program are point sources that must be included in the WLA portion of a TMDL" (USEPA 2002). Phase I and II permits can include the following types of discharges:

- Small, medium, and large municipal separate storm sewer systems (MS4s) – these can be owned by local jurisdictions, municipalities, and state and federal entities (e.g., departments of transportation, hospitals, military bases),

- Industrial facilities permitted for stormwater discharges, and

- Small and large construction sites

- Mineral mining facilities that do not have TSS limits

USEPA currently recommends that WLAs for NPDES regulated stormwater discharges be expressed as different WLAs for different identifiable categories (e.g., separate WLAs for MS4 and industrial stormwater discharges). These categories should be defined as narrowly as available information allows (e.g., for municipalities, separate WLAs for each municipality and for industrial sources, separate WLAs for different types of industrial stormwater sources or dischargers). In general, states are encouraged to disaggregate the WLA to facilitate implementation. USEPA recognizes that available data and information are usually not detailed enough to determine WLAs for NPDES regulated stormwater discharges on an outfall-specific basis (USEPA 2014).

The Upper Chester River NPDES Stormwater WLA is based on reductions applied to the sediment load from the portion of the urban land-use in the watershed associated with NPDES regulated stormwater permits. The NPDES stormwater WLA is calculated in the same manner as the load allocation for unregulated urban land acreage. Some of these sources may also be subject to controls from other management programs. The Upper Chester River NPDES Stormwater WLA does not require an overall reduction (see Table 8).

Table 8: Upper Chester River Watershed TMDL Reductions by Source Category

	Baseline Load Source Categories		Baseline Load (ton/yr)	TMDL Components	TMDL (ton/yr)	Reduction (%)
Upper Chester River Watershed contribution	**Nonpoint Source**		5,955	**LA**	4,732	21
	Point Source	Regulated Urban	15		15	0
		Individual Permits	10	**WLA**	10	0
Sub-total			**5,980**		**4,757**	**20**
Upstream contribution	Delaware Chester River		1,170	**Upstream LA**	916	22
Total			**7,150**		**5,673**	**21**

For more information on the methods used to calculate the NPDES regulated stormwater baseline sediment load, see Section 2.2.2. For a detailed list of all of the NPDES regulated stormwater discharges within the watershed and information regarding the NPDES stormwater WLA distribution amongst these discharges, please see the technical memorandum to this document entitled *Point Sources of Sediment in the Upper Chester River Watershed.*

As stormwater assessment and/or other program monitoring efforts result in a more refined source assessment, MDE reserves the right to revise the current NPDES Stormwater WLA provided the revisions protect aquatic life from sediment related impacts.

4.7 Upstream Load Allocation

The upstream loads from Delaware discharge into the non-tidal portion of the MD 8-digit Upper Chester River and are therefore included in this TMDL analysis. Maryland does not have any jurisdiction over upstream loads from Delaware, and therefore the load allocation presented is a recommendation and represents the reduction that will be required to meet the TMDL in Maryland. The Delaware portion of the Chester River is very similar in land use to the MD 8-digit. Therefore, the same reduction percentages were applied to non-regulated urban and crop loads. There are no wastewater or stormwater permits in the upstream Delaware Chester River.

4.8 Margin of Safety

All TMDLs must include a margin of safety (MOS) to account for any lack of knowledge and uncertainty concerning the relationship between loads and water quality (CFR 2012b). The MOS shall also account for any rounding errors generated in the various calculations used in the development of the TMDL. This TMDL was developed using an environmentally conservative approach that implicitly incorporates an MOS. Specifically, as was described in Section 4.2, the reference watershed forest normalized EOS loads were chosen in a conservative manner. Analysis of the reference group *forest normalized sediment loads* indicates that the 75th percentile of the reference watersheds is a value of 8.8 and that the median value is 5.9. For this analysis, the *sediment loading threshold* was set at the median value of 5.9 (MDE 2006). Use of the median as the threshold creates an environmentally conservative estimate, and results in an implicit MOS. Achieving a median forest normalized sediment load would assure that the watershed falls within the range of unimpaired watersheds.

4.9 Summary of Total Maximum Daily Loads

The average annual non-tidal Upper Chester River watershed TMDL is summarized in Table 9. The TMDL is the sum of the LA, NPDES Stormwater WLA, Wastewater WLA, and MOS. The LAs include nonpoint source loads generated within the Upper Chester River watershed and loads from upstream sources. The attainment of water quality standards within the non-tidal Upper Chester River watershed can only be achieved by

meeting the average annual TMDL of TSS specified for the watershed within this report. The Maximum Daily Load (MDL) is summarized in Table 10 (See Appendix B for more details).

Table 9: Upper Chester River Watershed Average Annual TMDL of TSS (ton/yr)

TMDL	=	LA				WLA				MOS
		LA_{DE}[1]	+	LA_{UCR}	+	NPDES Stormwater WLA_{UCR}	+	Wastewater WLA_{UCR}	+	
5,673	=	916	+	4,732	+	15	+	10	+	Implicit
		Upstream Load Allocations[1]			Upper Chester River Watershed TMDL Contribution					

Notes: [1] A delivery factor of 1 was used for all of the Upstream Load Allocations.

Table 10: Upper Chester River Watershed Maximum Daily Load of TSS (ton/day)

MDL	=	LA				WLA				MOS
		LA_{DE}	+	LA_{UCR}	+	NPDES Stormwater WLA_{UCR}	+	Wastewater WLA_{UCR}	+	
28.4	=	4.6	+	23.6	+	0.1	+	0.1	+	Implicit
		Upstream Load Allocations			Upper Chester River Watershed TMDL Contribution					

5.0 ASSURANCE OF IMPLEMENTATION

Section 303(d) of the CWA and current USEPA regulations require reasonable assurance that the sediment TMDL can and will be implemented (CFR 2012b). This section provides the basis for reasonable assurance that the sediment TMDL in the Upper Chester River watershed will be achieved and maintained.

While this TMDL establishes a sediment loading target for the watershed, watershed managers and other stakeholders should always remain cognizant that the endpoint of this TMDL, and hence the definition of its successful implementation, is based on in-stream biological health. Load reductions are critical to tracking this effort, since the TMDL target is defined as the point where sediment loads match those seen in reference watersheds, but the watershed cannot be delisted or classified as meeting water quality standards until it is demonstrated that the biological health of the stream system is no longer impaired by sediment. In planning any implementation efforts related to this TMDL, careful consideration should be given both to the sediment load reductions, and to the direct potential impacts on biological communities.

2010 Chesapeake Bay TMDLs
Implementation of the TMDL for sediment in the Upper Chester River watershed is expected to occur in parallel with implementation efforts for the 2010 Chesapeake Bay TMDLs for nutrients and sediment in the tidal fresh Upper Chester River. While the objectives of the two efforts differ, with the 2010 Bay TMDLs focused on tidal water quality and this TMDL targeting biological integrity in non-tidal streams, many of the sediment reductions achieved through implementation activities should result in progress toward both goals.

The strategies for implementing the 2010 Bay TMDLs are described in Maryland's Phase I WIP (MDE 2010) and Phase II WIP (MDE 2012). The WIPs are the centerpieces of the State's "reasonable assurance" of implementation for the 2010 Bay TMDLs, and the strategies encompass a host of BMPs, pollution controls and other actions for all source sectors that cumulatively will result in meeting the State's 2025 targets, as verified by the Chesapeake Bay Water Quality Sediment Transport Model. In particular, the implementation of practices to reduce sediment loadings from the agricultural and urban stormwater sectors should result in decreased loads in the Upper Chester River watershed's non-tidal streams.

Nonpoint Source Urban Lands
Generally speaking, urban areas that do not have NPDES permits do not have mandatory restoration requirements and restoration activities are largely voluntary. The State makes several efforts to encourage jurisdictions to conduct voluntary activities by providing technical assistance and funding opportunities to guide and support local actions. For example, Section 319 of the Clean Water Act provides federal grants to assist in nonpoint source (NPS) management. Section 319(b) requires preparation of a state NPS management program plan for approval by the US Environmental Protection Agency. Maryland's most recent plan, *Maryland's 2015-2019 Nonpoint Source Management*

Plan, addresses NPS according to pollutant and source. There are several programs listed in the report that address urban NPS, including Maryland Bay-Wise Program, Maryland Green Schools Awards, and the SMART Homeowner Reporting Program. Additionally, MDE is conducting outreach to non-MS4 jurisdictions regarding stormwater management requirements and retrofit BMPs. Funding sources for urban nonpoint source pollutants include: Federal 319(h) grants, Chesapeake and Atlantic Coastal Bays Trust Fund, and the State Revolving Loan Fund. More information on Maryland's NPS management program can be found at:
http://mde.maryland.gov/programs/Water/319NonPointSource/Pages/index.aspx.

Further efforts include offering competitive grant opportunities through the Chesapeake Bay Trust (CBT). Each year since 2015, MDE and DNR have entered into an agreement to pass federal funding through to local jurisdictions to enhance their ability to restore local water quality. The CBT administers the grant process with these funds and awards over $1.2M for use by regulated and non-NPDES jurisdictions. More information about this funding can be found on the CBT website here: https://cbtrust.org/wp-content/uploads/WAGP-2YR-Milestone-2017-2018_FINAL.pdf This grant opportunity combined with technical assistance offered to successful candidates is one more example of the State's plan to address the load allocations and to ensure a better level of reasonable assurance that the TMDL endpoints will be achieved.

Implementation of Agricultural Best Management Practices
In agricultural areas comprehensive soil conservation plans can be developed that meet criteria of the USDA-NRCS Field Office Technical Guide (USDA 1983). Soil conservation plans help control erosion by modifying cultural practices or structural practices. The reduction percentage attributed to cultural practices is determined based on changes in land-use, while structural practices have a reduction percentage of up to 25%. In addition, sediment loadings from livestock can be controlled via stream fencing and rotational grazing. Sediment reduction efficiencies of methods applicable to pasture land-use range from 40% to 75% (USEPA 2004). Lastly, riparian buffers can reduce the effect of agricultural sediment sources through trapping and filtering.

Maryland Funding Programs
In response to the WIP and the increased responsibility for local governments to achieve nutrient and sediment reduction goals, Maryland has continued to increase funding in the Chesapeake and Atlantic Coastal Bays Trust Fund. According to the Section 40 Report, even though the annual restoration funds for the four agencies [MDDNR, Maryland Department of Agriculture (MDA), MDE, Maryland Department of Planning (MDP)] varies from year to year, the total restoration funds for the first three years of the Chesapeake Bay WIP implementation evaluated time period (FY10 – FY12) was $882,327,165, while the total for the past four years of the period (FY12 – FY15) was $2,383,507,560, an increase of 170 percent. This increase was driven in part by the two primary Bay restoration Special Funds: The Bay Restoration Fund and the Chesapeake and Atlantic Coastal Bays Trust Fund (MDE et al. 2016). For more information on Maryland's implementation and funding strategies to achieve nutrient and sediment

reductions throughout the State's portion of the Chesapeake Bay watershed, please see [Maryland's Phase II Watershed Implementation Plan](#).

Some other examples of programs that can provide funding for local governments and agricultural sources include the Federal Nonpoint Source Management Program (§ 319 of the Clean Water Act), the Buffer Incentive Program (BIP), the State Water Quality Revolving Loan Fund and the Maryland Agricultural Water Quality Cost-Share Program.

In summary, through the use of the aforementioned funding mechanisms and BMPs, there is reasonable assurance that this TMDL can be implemented.

Additional Biological Stressors

As has been stated previously in this report, the biological impairment in this watershed is due to multiple stressors, not just sediment. While reducing TSS will bring about a water quality impact in terms of clarity, achieving a positive impact in stream biological communities might require several stressors to be addressed. These stressors were described in the Upper Chester River BSID report.

Many of the implementation actions to address sediment could concurrently address the other stressors identified in the BSID report. Since biological improvements will likely only be seen when multiple structural and pollutant stressors are addressed, watershed managers developing plans to address sediment should consider the effect of restoration projects on other stressors. Where possible, preference should be given to designs that address multiple stressors.

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APPENDIX A – Watershed Characterization Data

Table A-1: Reference Watersheds in the Coastal Plain Physiographic Region

MD 8-Name	MD 8-digit	Percent Stream Mile BIBI/FIBI < 3.0 (%)[1,2]	Forest Normalized Sediment Load[3]
Nassawango Creek	02130205	25	3.89
Tuckahoe Creek	02130405	18	5.88
Wye River	02130503	15	5.84
Langford Creek	02130506	18	11.13
Southeast Creek	02130508	0	5.87
Stillpond-Fairlee	02130611	22	9.75
Median			**5.9**
75th percentile			**8.8**

Notes: [1] Percent stream mile is based on the percentage of MBSS stations with BIBI and/or FIBI scores significantly lower than 3.0 within the watershed (MDE 2014b).

[2] The threshold to determine if an 8-digit watershed is impaired for impacts to biological   communities (IBI<3.0), is based on a comparison to reference conditions (MDE 2014b).

[3] Forest normalized sediment loads based on Maryland watershed area only (consistent with MBSS random monitoring data).

APPENDIX B – Technical Approach Used to Generate Maximum Daily Loads

Summary

This appendix documents the technical approach used to define maximum daily loads (MDLs) of sediment consistent with the average annual TMDL in the Upper Chester River watershed, which is considered the maximum allowable load the watershed can sustain and support aquatic life. The approach builds upon the modeling analysis that was conducted to determine the sediment loadings and can be summarized as follows.

- The approach defines MDLs for each of the source categories.

- The approach builds upon the TMDL modeling analysis that was conducted to ensure that average annual loading targets are at a level that support aquatic life.

- The approach converts daily time-series loadings into TMDL values in a manner that is consistent with available USEPA guidance on generating daily loads for TMDLs (USEPA 2007).

- The approach considers a daily load level of a resolution based on the specific data that exists for each source category.

Introduction

This appendix documents the development and application of the approach used to define MDL values. It is divided into sections discussing:

- Basis for approach

- Options considered

- Selected approach

- Results of approach

Basis for approach

The overall approach for the development of daily loads was based upon the following factors:

- **Average Annual TMDL:** The basis of the average annual sediment TMDL is that cumulative high sediment loading rates have negative impacts on the biological community. Thus, the average annual sediment load was calculated so as to ensure the support of aquatic life.

- **CBP P5.3.2 Watershed Model Sediment Loads:** As described in Section 2.2, the nonpoint source sediment loads from the Upper Chester River watershed are based on EOS loads from the CBP P5.3.2 watershed model. The CBP P5.3.2 model river segments were calibrated to daily monitoring information for

watersheds with a flow greater than 100 cubic feet per second (cfs), or an approximate area of 100 square miles.

- **Draft USEPA guidance document entitled "Developing Daily Loads for Load-based TMDLs":** This guidance document provides options for defining MDLs when using TMDL approaches that generate daily output (USEPA 2007).

The rationale for developing TMDLs expressed as *daily* loads was to accept the existing average annual TMDL, but then develop a method for converting this number to a MDL in a manner consistent with USEPA guidance and available information.

Options considered

The draft USEPA guidance document for developing daily loads does not specify a single approach that must be adhered to, but rather it contains a range of acceptable options (USEPA 2007). The selection of a specific method for translating a time-series of allowable loads into the expression of a TMDL requires decisions regarding both the level of resolution (e.g., single daily load for all conditions vs. loads that vary with environmental conditions) and level of probability associated with the TMDL.

This section describes the range of options that were considered when developing methods to calculate Upper Chester River watershed MDLs.

Level of Resolution

The level of resolution pertains to the amount of detail used in specifying the MDL. The draft USEPA guidance document on daily loads provides three categories of options for level of resolution, all of which are potentially applicable for the Upper Chester River watershed:

1. **Representative daily load:** In this option, a single daily load (or multiple representative daily loads) is specified that covers all time periods and environmental conditions.

2. **Variable daily load:** This option allows the MDL to vary as function of a particular characteristic that affects loading or waterbody response, such as flow or season.

Probability Level

All TMDLs have some probability of being exceeded, with the specific probability being either explicitly specified or implicitly assumed. This level of probability directly or indirectly reflects two separate phenomena:

1. Water quality criteria consist of components describing acceptable magnitude, duration, and frequency. The frequency component addresses how often conditions can allowably surpass the combined magnitude and duration components.

2. Pollutant loads, especially from wet weather sources, typically exhibit a large degree of variability over time. It is rarely practical to specify a "never to be exceeded value" for a daily load, as essentially any loading value has some finite probability of being exceeded.

The draft daily load guidance document states that the probability component of the MDL should be based on a representative statistical measure that is dependent upon the specific TMDL and the best professional judgment of the developers (USEPA 2007). This statistical measure represents how often the MDL is expected/allowed to be exceeded. The primary options for selecting this level of protection would be:

1. **The maximum daily load reflects some central tendency:** In this option, the MDL is based upon the mean or median value of the range of loads expected to occur. The variability in the actual loads is not addressed.

2. **The maximum daily load is a value that will be exceeded with a pre-defined probability:** In this option, a "reasonable" upper bound percentile is selected for the MDL based upon a characterization of the variability of daily loads. For example, selection of the 95th percentile value would result in a MDL that would be exceeded 5% of the time.

3. **The maximum daily load reflects a level of protection implicitly provided by the selection of some "critical" period**: In this option, the MDL is based upon the allowable load that is predicted to occur during some critical period examined during the analysis. The developer does not explicitly specify the probability of occurrence.

Selected Approach

The approach selected for defining an Upper Chester River Watershed MDL was based upon the specific data that exists for each source category. The approach consists of unique methods for each of the following categories of sources:

- Approach for Nonpoint Sources and Stormwater Point Sources within the Upper Chester River watershed

- Approach for Wastewater Point Sources within the Upper Chester River watershed

FINAL

<u>Approach for Nonpoint Sources and Stormwater Point Sources within the Upper Chester River Watershed</u>

The level of resolution selected for the Upper Chester River MDL was a representative daily load, expressed as a single daily load for each loading source. This approach was chosen based upon the specific data that exists for nonpoint sources and stormwater point sources within the Upper Chester River watershed. Currently, the best available data is the CBP P5.3.2 model daily time series calibrated to long-term average annual loads (per land-use). The CBP reach simulation results are calibrated to daily monitoring information for watershed segments with a flow typically greater than 100 cfs.

The probability level selected for the Upper Chester River MDL was a pre-defined exceedance probability. Based on the USEPA guidance, "in the case where a long term daily load dataset is available, in which multiple years of data and a variety of environmental conditions are represented, it is preferable to select a maximum daily load as a percentile of the load distribution. A sufficiently long-term dataset allows for minimizing error associated with the fact that the daily load dataset might not exactly match a normal or lognormal distribution" (USEPA 2007). The exact percentile value to be used should be determined by the TMDL developer, based on site specific characteristics.

This CBP P5.3.2 model output provides a time series of daily TSS loads from the Upper Chester River watershed, covering a 20-year period from 1985 to 2005. Because this is a long-term time series, it captures a broad range of meteorological and hydrological conditions and also minimizes the effect of potential statistical variances. As with the calculation of the TMDL value, environmentally conservative principles are also used in the MDL calculation. A 95th percentile flow was selected for the MDL, meaning that there is a 5% probability that daily loads will exceed this value. This percentile was chosen rather the 99th (which is also considered acceptable based on USEPA), in order to avoid the influence of extreme weather events and statistical outliers. Since the model daily time series represents the current (baseline) condition, the reduction percentage applied to each sector of the TMDL, was applied directly to the 95th percentile values to calculate the final MDL value.

$$MDL = 95th\ percentile\ of\ daily\ load\ series\ values * Reduction\ \%\ from\ TMDL$$

(Eq B-1)

Where:
   MDL = Maximum Daily Load, ton/day
   Daily load series values = CBP 5.3.2 output
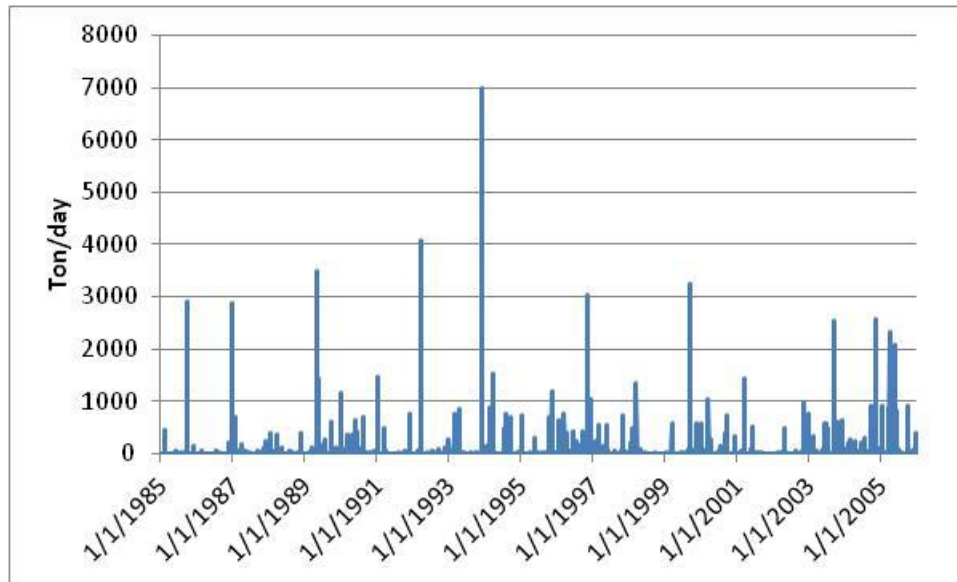   TMDL = Long term average annual load, ton/yr

Figure B-1: Daily Time Series of CBP River Segment Daily Simulation Results for the Upper Chester River Watershed

Approach for Wastewater Point Sources within the Upper Chester River Watershed

The TMDL also considers contributions from other point sources (i.e., sources other than stormwater point sources) in the watershed that have NPDES permits with sediment limits. As these sources are generally minor contributors to the overall sediment load, the TMDL analysis that defined the average annual TMDL did not propose any reductions for these sources and held each of them constant at their existing technology-based NPDES permit monthly (or daily if monthly was not specified) limit for the entire year.

The approach used to determine MDLs for these sources was dependent upon whether a maximum daily limit was specified within the permit. If a maximum daily limit was specified, then the reported average flow was multiplied by the daily maximum limit and a conversion factor of 0.0042 to obtain an MDL in ton/day. If a maximum daily limit was not specified, the MDLs were calculated based on the guidance provided in the Technical Support Document (TSD) for Water Quality-based Toxics Control (USEPA 1991). The long-term average annual TMDL was converted to maximum daily limits using Table 5-2 of the TSD assuming a coefficient of variation of 0.6 and a 99th percentile probability. This results in a dimensionless multiplication factor of 3.11. The average annual Upper Chester River TMDL of sediment/TSS is reported in ton/yr, and the conversion from ton/yr to a MDL in ton/day is 0.0085 (e.g. 3.11/365).

Results of approach

This section lists the results of the selected approach to define the Upper Chester River MDLs. The final results are presented in Table B-1.

- Calculation Approach for Nonpoint Sources and Stormwater Point Sources within the Upper Chester River Watershed

The MDL for Nonpoint Sources and Stormwater Point Sources within the Upper Chester River Watershed is based upon the 95th percentile value of the CBP P5.3.2 model daily load time series, reduced by the same percentage as the corresponding TMDL value. The 95th percentile load of the daily times series is 36 tons/day and with a TMDL reduction of 21%, it results in a total watershed MDL of 28.4 tons/day. The total MDL is subdivided in accordance with the same ratios present in the TMDL.

- Calculation Approach for Wastewater Point Sources within the Upper Chester River Watershed

    o For permits with a daily maximum limit:

Wastewater WLA_{UCR} (ton/day) = Permit flow (millions of gallons per day (MGD)) * Daily maximum permit limit (milligrams per liter (mg/l)) * 0.0042, where 0.0042 is a combined factor required to convert units to ton/day

    o For permits without a daily maximum limit:

Wastewater WLA_{UCR} (ton/day) = Average Annual TMDL Wastewater WLA_{UCR} Other (ton/yr)* 0.0085, where 0.0085 is the factor required to convert units to ton/day

The aggregate MDL for the point sources in the watershed is negligible.

Table B-1: Upper Chester River Watershed Maximum Daily Loads of Sediment/TSS (ton/day)

MDL	=	LA				WLA			+	MOS
		LA_{DE}	+	LA_{UCR}	+	NPDES Stormwater WLA_{UCR}	+	Wastewater WLA_{UCR}	+	
28.4	=	4.6	+	23.6	+	0.1	+	0.1	+	Implicit
		Upstream Load Allocations				Upper Chester River Watershed TMDL Contribution				